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## Unraveling assemblage of microbial community dwelling in Dabaoshan As/Pb/Zn mine-impacted area: A typical mountain mining area of South China

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1           Unraveling assemblage of microbial community dwelling in Dabaoshan  
2   As/Pb/Zn mine-impacted area: A typical mountain mining area of South China

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26 **Abstract**

27 Microbial community assemblage includes microorganisms from the three domains  
28 including Bacteria, Archaea, and Eukarya (Fungi), which play a crucial role in geochemical  
29 cycles of metal(loid)s in mine tailings. Mine tailings harbor vast proportions of metal(loid)s,  
30 representing a unique source of co-contamination of metal(loid)s that threaten the  
31 environment. The elucidation of the assembly patterns of microbial communities in  
32 mining-impacted ecospheres has received little attention. To decipher the microbial  
33 community assembly processes, the microbial communities from the five sites of the  
34 Dabaoshan mine-impacted area were profiled by the MiSeq sequencing of 16S rRNA  
35 (Bacteria and Archaea) genes and internal transcribed spacers (Fungi). Results indicated that  
36 the coexistence of 31 bacterial, 10 fungal, and 3 archaeal phyla, were mainly dominated by  
37 *Mucilaginibacter*, *Cladophialophora*, and *Candidatus Nitrosotalea*, respectively. The  
38 distribution of microorganisms was controlled by deterministic processes. The combination  
39 of Cu, Pb, and Sb was the main factor explaining the structure of microbial communities.  
40 Functional predicting analysis of bacteria and archaea based on the phylogenetic  
41 investigation of communities by reconstruction of unobserved states analyses revealed that  
42 the metabolic pathways related to arsenite transporter, arsenate reductase, and Fe-S cluster  
43 were important for metal detoxification. Furthermore, the ecological guilds (pathogens,  
44 symbiotrophs, and saprotrophs) of fungal communities explained 44.5% of functional  
45 prediction. In addition, metal-induced oxidative stress may be alleviated by antioxidant  
46 enzymes of fungi communities, such as catalase. Such information provides new insights into  
47 the microbial assembly rules in co-contaminated sites.

48 **Keywords:** Assembly process; Mine-impacted ecospheres; Co-occurrence network;  
49 Potential functional metabolism; Keystone taxa

50

## 51 **1. Introduction**

52 Sustainable environmental protection for non-ferrous mining wastes is vital for  
53 the development of the national economy, industry, science, and technology. The  
54 annual discharge of non-ferrous mining and smelting slag in the world exceeds 10  
55 billion tons, which are open-air stacked in mine tailing dams (Wang *et al.*, 2017). The  
56 content of metal(loid)s in non-ferrous tailings accounts for 30% - 90% of total  
57 industrial discharges, and flotation reagents were three times more enriched than  
58 pesticides in agricultural fields (Liu *et al.*, 2018; Zhang *et al.*, 2020). Such large  
59 quantities of tailings containing toxic, hardly decomposed, and complex multiple  
60 contaminants have seriously threatened food security and ecological security  
61 (Alakangas *et al.*, 2010; Liu *et al.*, 2022a; Zhu *et al.*, 2018). Complexation between  
62 metals and organic flotation reagents can accelerate the migration and transformation  
63 of metals, as well as inhibit the degradation of flotation reagents (Zhang *et al.*, 2020;  
64 Zhu *et al.*, 2018). Furthermore, exposure to toxicants in anthropogenic tailing dams by  
65 drinking the contaminated water, ingestions of soil and food grown on the waste, and  
66 dust inhalation could lead to dire consequences for human health (Hudson-Edwards,  
67 2016; Torres *et al.*, 2018; Wang *et al.*, 2017).

68 Arsenic (As) bearing minerals are ubiquitous in tailings, including sulfides,  
69 sulfosalts, arsenates, and arsenites (Giloteaux *et al.*, 2013). Arsenic has been listed as  
70 a suspected carcinogen by the International Agency for Research on Cancer, which is  
71 an agency of the World Health Organization. Additionally, the United States  
72 Environmental Protection Agency and the United Nations Environment Programme  
73 also recognize arsenic as a hazardous substance with potential carcinogenic properties.  
74 Arsenic, regarded as an environmental threat for all types of organisms, could induce  
75 cancers of the skin, lung, liver, and bladder among other tissues (Speer *et al.*, 2023).

76 Besides arsenic, elevated concentrations of associated metal(loid)s, such as copper  
77 (Cu), lead (Pb), and zinc (Zn), are frequently found in or around mine tailings  
78 worldwide (Hudson-Edwards, 2016).

79 Microorganisms (including Bacteria, Archaea, and Fungi) are important in the  
80 transformation of metal(loid)s ensuring the functions that sustain ecosystems. Bacteria  
81 are the dominant tailing-associated microorganisms (> 97%), followed by Archaea  
82 (0.26% - 2.75%), and Eukarya (0.21% - 0.11%) (Liu *et al.*, 2023). The distribution  
83 and composition of microorganisms can be driven by the spatial-temporal  
84 heterogeneity of geochemical factors, climate, and temperature (Liang *et al.*, 2020; Yi  
85 *et al.*, 2022). However, most studies have only focused on a single microbial  
86 component, i.e., bacteria (Liu *et al.*, 2019a; Sun *et al.*, 2018), fungi (Liu *et al.*, 2022b;  
87 Suting and Olivia Devi, 2021), or archaea (Chen *et al.*, 2021; Mendez-Garcia *et al.*,  
88 2015). Bacterial, fungal, and archaeal communities can contribute to the  
89 transformation of metal(loid) speciations and the degradation of flotation reagents in  
90 the co-contamination nonferrous mine environment. Intrinsic linkages between  
91 assembly processes and co-occurrences of bacteria, fungi, and archaea could facilitate  
92 the management of microorganisms for better provision of ecosystem services.  
93 However, to our knowledge, the co-occurrence of overall microbial assemblages  
94 including bacterial, fungal, and archaeal communities in tailings has not been well  
95 investigated.

96 To address this research gap, a study was conducted at a nonferrous mine tailing  
97 site with As/Pb/Zn co-contaminated in a Dabaoshan mine-impacted area (see  
98 Supplementary Information Fig. S1 and Table S1). We hypothesize here that: (1) the  
99 assembly processes of bacterial, fungal, and archaeal communities are influenced by  
100 various factors, including both biotic and abiotic elements, and (2) microbiological

101 communities with lower levels of species richness show greater linkages between  
102 species. We analyzed the samples collected from the Dabaoshan As/Pb/Zn  
103 mine-impacted area in a typical mountain mining area of South China. The objectives  
104 of this study were to (1) evaluate the diversity, structure, and composition of bacterial,  
105 fungal, and archaeal taxa in tailings; (2) reveal the ecological assembly processes and  
106 co-occurrence patterns of the three domain taxa under the long-term metal(loid)s  
107 co-contamination; and (3) explore the potential functional metabolism of taxa, which  
108 could be used for the management of microbial resources. This study provides new  
109 insights for studying community ecology, predicting responses to environmental  
110 changes, and designing strategies for microorganism management and site restoration.

## 111 **2. Materials and methods**

### 112 2.1 Site description, sampling, and physical-chemical analyses

113 The Dabaoshan mining area is located in the northern Guangdong province  
114 (southern China) and Lingnan mountain region with a hilly landform (24°30'01" -  
115 24°35'26" N, 113°41'53" - 113°46'40" E, Fig. S1). The Dabaoshan mine, a typical  
116 mountainous mining area, is the largest in South China (Wang *et al.*, 2019). It is a  
117 large polymetallic mine composed mainly of Fe and Cu. It has been mined  
118 extensively since the 1970s, and most of the metal ores are S-containing ores such as  
119 pyrite, pyrrhotite, and chalcopyrite, which are important for the generation of acid  
120 mine drainage (Bao *et al.*, 2018). The upper part of the polymetallic deposit was a  
121 weathered and leached limonite, the middle part was layered siderite, and the lower  
122 part was Cu, Pb, Zn, Fe, and S deposits, and porphyry Mb and Mb polymetallic  
123 deposits (Luo *et al.*, 2021). The tailings site had been used for open storage of Pb-Zn  
124 or Fe smelting slag for almost 30 years and was abandoned for almost 20 years. The  
125 total amount of various solid wastes is over 9 million tons.

126 The sampling area covers a total surface area of 0.47 ha (Fig. S1). This area was  
127 selected based on previous physical-chemical characterization studies (Zhang *et al.*,  
128 2019), and has been undergoing natural attenuation for almost 20 years. The sampling  
129 area was divided into five general sub-areas, which were divided along the directions  
130 of the surrounding LiWu river (Fig. S1) and showed different physical-chemical  
131 properties characteristics (Table S1). Due to the irregular terrain with high  
132 heterogeneity of tailings, a random sampling strategy was adopted collecting  
133 composite samples by nested sampling within a site. Composite samples using a  
134 nested sampling design within a site accounted for environmental heterogeneity and  
135 were reasonably reliable for the evaluation of the microbial diversity variation at a  
136 small scale (Liu *et al.*, 2019).

137 For each sub-area, 13 samples were collected in December 2021 based on a  
138 squared-figure sampling design. Samples were then combined and homogenized to  
139 produce a composite sample to evaluate variations of bacteria, fungi, and archaea  
140 communities. A total of five composite samples were finally harvested after field  
141 sampling. All composite samples were immediately kept on ice (at 4°C) and delivered  
142 to the China University of Geosciences (Beijing) within 48 h of sampling. Samples  
143 for physical-chemical analysis were air-dried and sieved at 100-mesh size (0.149 mm,  
144 US standard).

145 The detection of pH was done following the procedure of the China National  
146 Agricultural Technology Extension Center (2006). After digestion in a microwave  
147 with HNO<sub>3</sub>-HCl-HF (5:3:2, v/v/v, GR), the concentrations of metal(loid)s were  
148 measured by inductively coupled plasma-mass spectrometry (ICP-MS). More detailed  
149 information for the measurements of pH and metal(loid)s concentrations was  
150 previously described by Liu *et al.* (2021). The mineralogical composition of the

151 sample was determined using a scanning electron microscope (SEM) and energy  
152 dispersive spectroscopy (EDS) spectrometer analysis, combined with phase analysis  
153 of X-ray diffraction (XRD) and X-ray fluorescence (XRF). For microorganism  
154 analysis, samples from each site were stored at -80 °C for further DNA extraction and  
155 Miseq sequencing.

## 156 2.2 Amplification, MiSeq sequencing, and data processing

157 Total genomic DNA was extracted from ~ 10.0 g of each tailing sample using a  
158 FastDNA® spin kit for soil (MP, Biomedicals, USA). The extracted DNA was  
159 subjected to PCR amplification using the primer sets of: 1) ArBa515F  
160 (GTGCCAGCMGCCGCGGTAA) and 806R (GGACTACHVGGGTWTCTAAT),  
161 targeting the V3-V4 region of bacterial 16S rRNA gene (Liu *et al.*, 2023); 2) ITS1-F  
162 (CTTGGTCATTTAGAGGAAGTAA) and ITS2-R  
163 (GCTGCGTTCTTCATCGATGC), targeting the fungal genomic ITS region (Liu *et*  
164 *al.*, 2022c); 3) 524F10extF (TGYCAGCCGCGCGGTAA) and Arch958RmodR  
165 (YCCGGCGTTGAVTCCAATT), targeting the V3-V4 region of archaeal genomic  
166 gene (Yi *et al.*, 2019). Polymerase chain reaction (PCR) reaction (20 µL) was carried  
167 out containing 10 µL of 2×Pro Taq, 0.8 µL of forward/reverse primers (5 µM), 10  
168 ng/µL of extracted template DNA, and double-distilled water. The PCR reaction was  
169 performed on an ABI GeneAmp® 9700 Thermal Cycler with the procedures: initial  
170 denaturation (3 min at 95°C), followed by 27 cycles of 30 s at 95°C, annealing 30 s  
171 (at 55°C for bacteria, 55°C for archaea, and 50°C for fungi), and extension (45 s at  
172 72°C), and a final extension (10 min at 72°C). This reaction was carried out (in  
173 triplicate) for each composite sample, and then these samples were mixed into a new  
174 PCR composite sample for sequencing. Sequencing was conducted on an Illumina  
175 MiSeq platform at Majorbio Bio-Pharm Technology Co., Ltd (Shanghai, China).



176 The acquired sequences were filtered, and chimeric sequences were removed  
177 using the USEARCH7-uparse clustering algorithm. Sequences were then split into  
178 operational taxonomic units (OTUs) at a 97% similarity level according to 70%  
179 classification confidence as previously described (Mamet *et al.*, 2017; Liu *et al.*,  
180 2019a). All raw data were trimmed and filtered using fastp software (v0.19.4) by  
181 filtering the reads below 50bp and reads with nitrogenous base. The fast length  
182 adjustment of short reads (FLASH) software was used for merging the paired-end  
183 reads of raw sequences. Sequences of bacterial/archaeal, and fungal were assigned to  
184 taxonomic based on the taxonomy database of silva138/16s and unite8.0/its\_fungi,  
185 respectively. The sequence read archive accession number for the sequences is  
186 SUB13926409.

#### 187 2.4 Statistical analyses

188 To determine the differences in alpha-diversities of bacterial, fungal, and  
189 archaeal communities, analyses based on the Wilcoxon test were calculated using  
190 OTUs tables resampled to a minimum number of sequences from each sample  
191 (39,922 for bacteria, 72,629 for fungi, and 42,962 for archaea). In present study,  $p <$   
192 0.05 is considered statistically significant. To estimate the beta-diversity of bacterial,  
193 fungal, and archaeal communities, non-metric multidimensional scaling (NMDS) was  
194 conducted based on the Bray-Curtis dissimilarity algorithms. The Partial Mantel test  
195 determined the correlations between compositional and environmental data based on  
196 Euclidean distances (Sunagawa *et al.*, 2015). BIOENV analysis was used to determine  
197 the combined effects of physical-chemical parameters on the distribution of microbial  
198 communities. The incidence-based (Raup-Crick) dissimilarity indices ( $\beta_{RC}$ ) were  
199 computed to determine whether communities were stochastically or deterministically  
200 assembled. The neutral community model was constructed to evaluate the

201 contribution of neutral processes in community structure as described by others  
202 (Sloan *et al.*, 2006). The Levins' niche breadth index for meta-communities was  
203 estimated to reveal the patterns of meta-communities sorting and dispersal limitation  
204 and their adaptability to the contaminated tailings. A high breadth index for a given  
205 community indicates its wide habitat niche breadth. The wider the species niche, the  
206 species is less specialized and more metabolically flexible. Co-occurrence network  
207 analyses of bacterial, fungal, and archaeal communities were conducted by Gephi  
208 software (v0.9.2) using the Frucherman Reingold algorithm as described by others  
209 (Han *et al.*, 2017). The degree of the node of the co-occurrence network was assessed  
210 by the number of nodes connected directly to that node. The more lines on the node,  
211 the higher the degree, which indicates the intensity of correlation between the species  
212 and other species.

213 To predict the potential functions of bacterial and archaeal communities, the  
214 Phylogenetic Investigation of Communities by Reconstruction of Unobserved States  
215 (PICRUSt2) was used to reveal the functions encoded in communities based on the  
216 MetaCyc database. The weighted nearest sequenced taxon index (NSTI) was  
217 calculated to assess the accuracy of PICRUSt2 analysis (Langille *et al.*, 2013). Kyoto  
218 Encyclopedia of Genes and Genomes (KEGG) databases (e-value cut-off  $10^{-5}$ ) were  
219 used for functional annotation. Significant differences in KOs between bacterial and  
220 archaeal communities were analyzed based on the independent samples *T*-test. To  
221 identify different functional groups within fungal communities, a recently developed  
222 open annotation tool (FunGuild, Fungi Functional Guild) was used for the assignment  
223 of fungal communities (Nguyen *et al.*, 2016). Redundancy analysis was used to  
224 examine the correlations between KOs/Guilds and physical-chemical parameters  
225 based on a weighted normalized unifracs distance algorithm. All the analyses were

226 done using R software (v 3.4.1) unless otherwise stated.

227

### 228 **3. Results and discussion**

#### 229 3.1 The morphological characteristics of main mineral grains

230 The pH in the five studied samples ranged from 4.03 to 6.56 (Table S1), being  
231 higher than the reported abandoned tailings (Liu *et al.*, 2019b; Liu *et al.*, 2021).

232 Generally, the heavy mineral fractions have typical metal(loid) concentrations that are  
233 one order of magnitude greater, indicating that the majority of the contaminants are  
234 likely bonded to high-density phases (Tuhý *et al.*, 2020). The metal(loid)  
235 concentration (mg/kg) in the mineral fractions were in the following ranges: As (from  
236 2180 to 2810), Fe (72900 - 97900), Pb (2800 - 4060), and Zn (1260 – 2220) (Table  
237 S1). The levels of As, Fe, Pb, and Zn were from 1.36 to 17.98 times higher than the  
238 soil environmental quality standard (GB15628-1995 and GB36600-2018).

239 Presumably, this contamination followed the release of metal(loid)s from smelter  
240 tailing after > 30 years of abandonment (Alakangas *et al.*, 2010). According to the  
241 ICP-MS analysis, the major potential toxicants were As, Pb, and Zn, resulting in  
242 metal(loid) pollution in the Dabaoshan mine-impacted area. Correlation analysis  
243 between metal(loid)s showed that As, Pb, Sb, and Zn had a significant positive  
244 correlation with each other ( $r = 0.878 - 0.978$ ;  $p < 0.05$  or  $p < 0.01$ ; Table S2) in the  
245 studied sites. As well, Cr had a high correlation coefficient with Cu ( $r = 0.916$ ;  $p <$   
246  $0.05$ ; Table S2). Results suggested that the elements had the same source, with similar  
247 physical-chemical properties.

248 The XRD analysis indicated that samples were predominantly composed of quartz,  
249 manganese copper oxide, Iwakiite, Celadonite, and Magnesium iron oxide (Fig. 1).

250 These five minerals, except quartz and Iwakiite, were often observed in the gangue  
251 minerals. Some very complex minerals were also present as other minor phase forms  
252 (Fig. 1). The presence of minor mineral phases may correspond to the  
253 high-temperature amorphous and nanocrystalline phases produced during the smelting  
254 and refining operations (Tuhý *et al.*, 2020; Zhu *et al.*, 2021). Representative  
255 SEM-EDS images of the typical minerals are shown in Figure 1.

256 < Insert Fig. 1 >

257 The main minerals presented as irregular large particles and were composed of  
258 various mineral grain aggregates. Moreover, the elemental composition (%) of tailings  
259 samples consisted mainly of C (from 5.35 to 33.85), O (40.33 - 61.91), P (0.28 - 1.87),  
260 Si (5.00 - 21.94), and S (0.15 - 2.11) (Table S3). The major metals were Al (6.21 -  
261 17.84), As (0.11 - 1.36), Fe (1.00 - 38.96), Pb (0.16 - 4.47), and Zn (0.09 - 1.61), as  
262 confirmed by XRF analysis (Table S4). It is likely that As and Zn participated in the  
263 silicate structure (Zhu *et al.*, 2021). These results differed from data from samples  
264 collected from Namibian mining/smelting-polluted soils (Tuhý *et al.*, 2020). Such  
265 different mineral structures are probably due to different mining/smelting processes  
266 adapted to the ore grade and the target extracted metals (Alakangas *et al.*, 2010; Liu *et*  
267 *al.*, 2019a,2020; Tuhý *et al.*, 2020).

### 268 3.2 Characteristics of microbial communities

269 A total of 31 bacterial (1042 OTUs), 10 fungal (328 OTUs), and 3 archaeal (70  
270 OTUs) phyla were obtained, which are consistent with earlier metagenomic analyses  
271 showing that ~ 97% of microorganisms in tailings are assigned to bacteria (Liu *et al.*,  
272 2023). The microbial diversity, relative abundance, and phylogenetic diversity  
273 following down-weighting changed as follows: bacteria, fungi, and archaea (Fig. 2a).

274 The boneh and coverage estimates indicated that there would be no additional  
275 observed OTUs for an additional level of sampling (coverage index > 99.87%). There  
276 were differences in alpha-diversity indices (such as Shannon, Simpson, Chao1, ace,  
277 and coverage) among bacterial, fungal, and archaeal communities ( $p < 0.05$ ). The  
278 distributions of OTUs based on BIOENV analysis appeared strongly driven by the  
279 combination of pH, Pb, and Zn (bacteria OTUs,  $r = 0.77$ ), Fe and Pb (fungi OTUs,  $r =$   
280  $0.77$ ), and Fe and Pb (archaea OTUs,  $r = 0.37$ ) (Table S5). These results suggested  
281 that physical-chemical parameters influenced the distribution of the OTUs from the  
282 different microbial domains, which was consistent with other studies showing that the  
283 dynamics of physical-chemical parameters as well as their complex combinations can  
284 drive the microbial structure (Mamet *et al.*, 2017; Yi *et al.*, 2022).

285 < Insert Fig. 2 >

286 Among the 31 bacterial phyla, *Pseudomonadota* (25.3% - 38.6%), *Acidobacteriota*  
287 (0.62% - 18.3%), and *Nitrospirota* (0.16% - 23.9%) dominated the tailings sites (Fig.  
288 2b). *Pseudomonadota* and *Acidobacteriota* are common phyla in tailings from  
289 different locations (Liu *et al.*, 2018). *Nitrospirota* includes only three valid and  
290 published genera, harboring potentially dissimilatory sulfur metabolisms (Umezawa  
291 *et al.*, 2020). In the fungal communities, *Ascomycota* and *Basidiomycota* accounted  
292 for > 95.2% of the sequences across all tailings (Fig. 2c). These fungal phyla are  
293 dominant in various environments, such as soil (Moussa *et al.*, 2017) and lead-zinc  
294 tailings (Jia *et al.*, 2021). Archaeal communities were dominated by  
295 *Thermoplasmatota* (56.6%) and *Crenarchaeota* (43.3%) phyla (Fig. 2d). Evidence  
296 showed that the *Thermoplasmatota* accounted for 75.1%, while *Crenarchaeota* were  
297 not represented in an acid mine drainage environment of France (Bruneel *et al.*, 2008;  
298 Calum 2023). This is likely due to the different geological environment conditions

299 and tailings types. *Thermoplasmatota* and *Crenarchaeota*, which harbor aerobic  
300 respiration and acid tolerance genes, could contribute to the indigenous microbial  
301 genetic repertoire to expand into a new habitat (Brochier-Armanet *et al.*, 2008;  
302 Sheridan *et al.*, 2022). Also, there are few studies on the composition and structure of  
303 the archaeal communities in metal(loid) tailings. In the present study, most of the  
304 sequences could not be identified for bacteria (54.2%), fungi (26.1%), and archaea  
305 (44.4%) (Fig. 2b-d). The microbial identification described here was lower than that  
306 reported for other tailings such as Sb-rich or multi-contaminated nonferrous  
307 metal(loid) tailings dumps (Liu *et al.*, 2019b; Xiao *et al.*, 2016). Such observation  
308 suggests that a vast microbial population in tailings remains under-exploited; their  
309 exploration would benefit bioremediation and environmental management (Liu *et al.*,  
310 2019a).

### 311 3.3 Potential processes governing the microbial assemblage

312 We also explored the microorganism taxonomic classifications and their relative  
313 distributions in tailings. The NMDS ordination analysis based on Bray-Curtis  
314 similarity distance assessed the beta-diversity of the bacterial, fungal, archaeal, and all  
315 microbial communities in tailings (Fig. 3a-d). Bacterial and fungal communities were  
316 only detected in four sampling sites except at the T2 site (Fig. 2b,c), and tailing sites  
317 (T3 and T5) with a pH > 5 were compared to that < pH 5 (T1 and T4) (Fig. 3a,b).

318 < Insert Fig. 3 >

319 The pH plays a critical role in affecting the abundance and distribution of microbes  
320 (Chen *et al.*, 2014). Unexpectedly, archaeal communities were detected in five sites,  
321 including the T2 sampling site (Fig. 2d), which was distant from the other sites (Fig.  
322 3c). Also, the distribution of bacterial and fungal communities was more similar to

323 each other than that of archaeal communities ( $p = 0.018$ , stress = 0.073, permutations  
324 = 999; Fig. 3d). The environmental factors driving the relationships between the three  
325 domains of life were examined based on Mantel test (Fig. 3b). Overall, the Cu, Pb,  
326 and Sb contents were the major factors correlating with the abundance distribution of  
327 bacteria, fungi, and archaea ( $p < 0.05$ ; Fig. 3b), whereas the pH, Fe, and Zn contents  
328 were not correlated ( $p > 0.05$ ; Fig. 3b). Such observation probably reflects the  
329 dissimilarity distribution of OTUs among the sites, suggesting that the microbial  
330 composition exhibits habitat-specific patterns. This confirms the possibility that the  
331 distribution and composition of microbial communities are changed by dynamic and  
332 rapidly changing conditions, likely caused by anthropogenic activity-induced stresses,  
333 temperature, oxygen, salinity, latitude, and seasonal changes (Cabral *et al.*, 2016;  
334 Cravo-Laureau *et al.*, 2017; Giloteaux *et al.*, 2013; Pringault *et al.*, 2008).  
335 Nevertheless, in natural environments, the composition and distribution of  
336 microorganisms interact with both biotic and abiotic components (Cravo-Laureau *et*  
337 *al.*, 2017). Network analysis showing the co-occurrence abundance of members of the  
338 three domains of life indicated that *Mucilaginibacter*, *Cladophialophora*, and  
339 *Candidatus Nitrosotalea* were found as Specialists, being the main contributing  
340 genera (Fig. 4).

341 < Insert Fig. 4 >

342 It is also worth mentioning that co-occurrence analysis identified the rare taxa of  
343 *Mucilaginibacter* and *Cladophialophora* (< 1.79%; Fig. 2b,c) identified as Specialist,  
344 corresponding to keystone taxa organizing the microbial assemblage in tailings.  
345 Recent studies explored that keystone taxa have a pivotal role in driving microbial  
346 structure and functioning (Banerjee *et al.*, 2018; Herren and McMahon, 2018). Both  
347 *Mucilaginibacter* and *Cladophialophora* contain members resistant to potential toxic

348 metal(loid)s and can adapt to contaminated environments (Taleski *et al.*, 2020;  
349 Vasconcelos *et al.*, 2021). Additionally, the dominated archaeal members of  
350 *Candidatus Nitrosotalea* (16.2% - 70.6%; Fig. 2d) were also affiliated with keystone  
351 taxa. *Candidatus Nitrosotalea*, an obligate acidophilic ammonia oxidizer, contains  
352 genes encoding predicted potential pH homeostasis mechanisms (Lehtovirta-Morley  
353 *et al.*, 2016). Thus, our results revealed that both rare and abundant taxa contribute to  
354 ecosystem stability and function, which is consistent with previous reports in coastal  
355 systems (Hugoni *et al.*, 2013).

356 To define the relative importance of deterministic and stochastic processes in  
357 shaping microbial community structure incidence-based (Raup-Crick) beta-diversity  
358 ( $\beta_{RC}$ ) tests were conducted (Fig. 5a). The mean values of  $\beta_{RC}$  were close to |1| for the  
359 three domains of life, indicating deterministic processes control microbial  
360 assemblages. For the bacterial communities, the deterministic processes favored  
361 dissimilar communities, whereas for archaeal and fungal communities, the  
362 deterministic assemblage processes produced more similar communities than  
363 expected by chance. However, neutral community model analysis revealed a  
364 negligible degree of stochastic processes (Fig. 5b and Fig. S2), confirming its minor  
365 role in the assembly of microbial communities. In addition, fungal communities  
366 demonstrated wider niche breadths than the archaeal and bacterial communities (Fig.  
367 5c), indicating higher fungal metabolic flexibility and redundancy. A possible  
368 explanation is that the Dabaoshan mine-impacted area is a unique habitat due to  
369 long-term abandonment. This As, Pb, and Zn co-contamination resulted in specific  
370 environmental filtering that influences the niche breadth such as abiotic conditions,  
371 biotic interactions, or geographical ranges (Malard *et al.*, 2021).

372 < Insert Fig. 5 >



### 373 3.4 Potential metabolism and predicted functions

#### 374 3.4.1 Metal-related functions of bacterial and archaeal communities

375 The bacterial and archaeal metabolic potentials were inferred by PICRUSt2  
376 analysis (Fig. S3). The obtained NSTI values for both bacterial and archaeal  
377 communities were low (0.00 - 0.36; Table S6), and the mean fraction of sequences  
378 used in PICRUSt2 prediction were > 99.89% (Table S6), indicating that the prediction  
379 was accurate (Douglas *et al.*, 2019; Langille *et al.*, 2013). The extremely high  
380 concentrations of As, Fe, Pb, and Zn in the studied tailings likely enriched the  
381 abundance of metal-resistance genes during the long term (Table S1) compared to  
382 short-term exposures (Liu *et al.*, 2021). Among the 6779 (for bacterial community)  
383 and 3145 (for archaeal community) KOs, 23 metal-related KOs were shared for  
384 bacterial and archaeal communities (Fig. S3; Table S7 and S8). Relative abundances  
385 of KOs were related to arsenite transporter (K03325), arsenate reductase (K03741),  
386 and Fe-S cluster assembly proteins (K09013-9015) differed between bacterial and  
387 archaeal communities ( $p < 0.05$ ; Fig. 6a).

388 < Insert Fig. 6 >

389 It was noteworthy that the detected arsenate reductase and arsenite transporter could  
390 reduce  $\text{As}^{5+}$  to  $\text{As}^{3+}$  to be retained in the bacterial cell, released into the air, or both  
391 (Rahman *et al.*, 2021). Although the expression of KO genes related to arsenate  
392 respiratory reductase (*arr*) may not be strictly anaerobic (Jia *et al.*, 2014), the KOs  
393 were not revealed by PICRUSt2 here (Fig. S3). Microorganisms harboring *arr* genes  
394 can use arsenate as a terminal electron acceptor for respiration in cell periplasm  
395 (Bhattacharjee and Rosen, 2000). Thus, a proportion of microorganisms should be  
396 active for arsenic detoxification in the studied sites. The KOs related to Fe-S cluster

397 assembly protein were dominant for archaeal communities. The Fe-S cluster assembly  
398 could traffic the toxic iron and sulfide to the target metalloproteins to maintain the  
399 Fe-S homeostasis (Chahal *et al.*, 2009). The shared KOs related to Co-Zn-Cd efflux  
400 system protein (K16264) and arsenite/tail-anchored protein-transporting ATPase  
401 (K01551) were also detected (Tables S7 and S8). The arsenite/tail-anchored  
402 protein-transporting ATPase could couple with arsenic pump membrane protein for  
403 the cellular expulsion of arsenite or antimonite (Bhattacharjee and Rosen, 2000).  
404 Additionally, specific KOs between bacterial and archaeal communities were also  
405 detected (Fig. S3 and Table S7). The distinct distribution of these predicted functional  
406 pathways was strongly correlated with As, Fe, and Pb contents ( $r = 0.98$ ; Fig. 6b,c).  
407 Additional contamination, whether organic or metallic/metalloid, could modify the  
408 composition and function of microorganisms in natural habitats (Ning *et al.*, 2020;  
409 Mendez-Garcia *et al.*, 2015). For bacterial taxa, specific KOs related to the  
410 arsenate/arsenite/antimonite responsive transcriptional repressor (K03892), F-type  
411 H<sup>+</sup>-transporting ATPase subunit *b* (K02108 - K02115), Fe<sup>3+</sup> transporters (K02010,  
412 K02011, and K02012), and metal efflux systems (K15726, K07787) showed high  
413 relative abundance (Fig. S3a). The KOs related to Fe<sup>3+</sup> transport and metal efflux  
414 could mediate resistance to Co<sup>2+</sup>, Zn<sup>2+</sup>, Cd<sup>2+</sup>, and Fe<sup>3+</sup> and function as a cation/proton  
415 antiport for the removal of cellular cations (Cabral *et al.*, 2016). Given the presence of  
416 several metal-efflux pumps and metal-regulated gene expression in tailings, these  
417 processes may be crucial in the detoxification of metals (Maynaud *et al.*, 2014; Ren *et*  
418 *al.*, 2017). For archaeal taxa, the dominant KOs related to arsenite/tail-anchored  
419 protein-transporting ATPase (K01551) were specific to the Dabaoshan As/Pb/Zn  
420 mine-impacted area (Fig. S3b and Table S8). To adapt to the extreme tailings  
421 environment, microorganisms harbored specific functions known as essential for the

422 biogeochemical cycling and biodegradation of pollutants (Cravo-Laureau *et al.*,  
423 2017).

#### 424 3.4.2 Functional prediction of fungal communities

425 We assigned fungal OTUs into specific ecological guilds based on FUNGuild  
426 (Fig. S4a). Three ecological guilds (pathogens, symbiotrophs, and saprotrophs)  
427 accounted for ~ 44.5% of all the fungal OTUs (Fig. S4a). It was surprising to detect  
428 relatively high numbers of animal pathogens and soil saprotrophs in the studied  
429 tailings (Fig. S4a). The animal pathogens always showed a high abundance in the  
430 forest or soil system with different degrees of residency: permanent, periodic,  
431 transient, and incidental (Samaddar *et al.*, 2021). The relatively high abundance of  
432 animal pathogens seen here was likely introduced anthropogenically or after surface  
433 runoff or aerial disposition. Soil saprotrophs were the key fungi in the decomposition  
434 of recalcitrant organic C, conducting biogeochemical cycles of key nutrients (Liu *et*  
435 *al.*, 2022c). Correlation analyses using the geochemical parameters showed that the  
436 distribution of functional fungal guilds, such as soil saprotroph was correlated with  
437 concentrations of As, Fe, and Pb ( $p < 0.05$ ; Fig. S4c). Such observation further  
438 supports that deterministic processes are the most important process driving fungal  
439 distributions in tailings. Also, the newly updated PICRUST2 software was utilized to  
440 explore the functional features of fungal communities (Fig. S4b). The relevant  
441 antioxidant enzymes, such as catalase, glucose-6-phosphate dehydrogenase,  
442 glutathione reductase, peroxidase, and superoxide dismutase were selected from the  
443 functional predictions (Fig. S4b and Table S9). These metal antioxidant enzymes can  
444 mitigate the membrane lipid peroxidation damage caused by metal stress, and  
445 maintain the community stability (Azevedo *et al.*, 2007; Garg and Kaur, 2013). These  
446 antioxidant enzymes could enhance the coping mechanism to metal-induced oxidative

447 stress in As/Pb/Zn contamination tailings. Thus, the functional fungal taxa adapting to  
448 metal stress may be potential candidates for tailings restoration. Restoration based on  
449 indigenous sulfate-reducing bacteria consortium has been applied to the maintenance  
450 and stabilization of metal(loid)s in tailings (Liu *et al.*, 2019a; Li *et al.*, 2022).  
451 Microorganisms inhabiting mine tailings require specific metabolic strategies for their  
452 survival and may hold potential resources for future pollution cleanup strategies.  
453 Effective in situ bioremediation will rely on an in-depth understanding of the function  
454 of the bacterial communities, especially the abundant and metabolically active  
455 phylotypes.

## 456 **5. Conclusions**

457 The present study provides new insight into the co-occurrence patterns of  
458 bacteria, fungi, and archaea communities in co-contaminated typical metal(loid)  
459 tailings after 20 years of abandonment. The combined mineralogy analyses revealed  
460 that As, Pb, and Zn were the major toxicants. The microbial distribution patterns  
461 indicated that the assemblages of bacterial, fungal, and archaeal taxa were influenced  
462 by the metal(loid)s (including Cu, Pb, and Sb), which suggests that deterministic  
463 processes play a key role in shaping microbial community assemblages. The  
464 microbial community arranges around keystone taxa such as members of  
465 *Mucilaginibacter*, *Cladophialophora*, and *Candidatus Nitrosotalea* genera. The  
466 PICRUSt analysis showed that specific functional KOs attested to the distinct roles of  
467 bacteria and archaea communities prevailing in the metal detoxification of tailings,  
468 such as As<sup>5+</sup>. The relevant antioxidant enzymes of fungi taxa (catalase and glutathione  
469 reductase) could relieve metal-induced oxidative stress. To the best of our knowledge,  
470 the present study provides, for the first time, new insight into how microbial  
471 communities (including Bacteria, Archaea, and Fungi) respond to typical As/Pb/Zn

472 mine tailings. Such emerging technologies are applicable for the future restoration of  
473 soils co-contaminated with metal(loid)s from mine tailings.

474

#### 475 **Credit author statement**

476 **Jian-li Liu**: Samples collecting, Data Analysis, Writing - original draft; **Jun Yao**:  
477 Writing-Reviewing and Editing, and Validation; **Ruofei Li**: Samples collecting;  
478 **Houquan Liu** and **Jun-jie Zhu**: Morphological analysis of main mineral grains;  
479 **Geoffrey Sunahara** and **Robert Duran**: Writing - Review & Editing.

480

#### 481 **Declaration of competing interest**

482 The authors declare that no conflict of financial interest or personal relationships  
483 exists in the submission of this manuscript.

#### 484 **Data availability**

485 Data will be made available on request.

486

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493

#### 494 **Appendix A. Supporting information**

495 Supplementary data related to this article can be found on the website of *Science of*  
496 *the Total Environment*.

497

498 **References**

- 499 Alakangas L., Ohlander B., Lundberg A., 2010. Estimation of temporal changes in oxidation  
500 rates of sulphides in copper mine tailings at Laver, Northern Sweden. *Sci. Total Environ.*  
501 408, 1386-1392. <https://doi.org/10.1016/j.scitotenv.2009.11.005>
- 502 Azevedo M.M., Carvalho A., Pascoal C., Rodrigues F., Cássio F., 2007. Responses of  
503 antioxidant defenses to Cu and Zn stress in two aquatic fungi. *Sci Total Environ.*  
504 377(2-3), 233-43. <https://doi.org/10.1016/j.scitotenv.2007.02.027>
- 505 Banerjee S., Schlaeppli K., van der Heijden, M.G.A., 2018. Keystone taxa as drivers of  
506 microbiome structure and functioning. *Nat. Rev. Microbiol.* 16, 567-576.  
507 <https://doi.org/10.1038/s41579-018-0024-1>
- 508 Bao Y., Guo C., Lu G., Yi X., Wang H., Dang Z., 2018. Role of microbial activity in Fe(III)  
509 hydroxysulfate mineral transformations in an acid mine drainage-impacted site from the  
510 Dabaoshan Mine. *Sci Total Environ.* 616-617, 647-657.  
511 <https://doi.org/10.1016/j.scitotenv.2017.10.273>
- 512 Bhattacharjee H., Rosen B.P., 2000. Role of conserved histidine residues in metalloactivation  
513 of the *ArsA* ATPase. *Biometals* 13, 281-288. <https://doi.org/10.1023/a:1009200215328>
- 514 Brochier-Armanet C., Boussau B., Gribaldo S. Forterre P., 2008. Mesophilic crenarchaeota:  
515 Proposal for a third archaeal phylum, the *Thaumarchaeota*. *Nat. Rev. Microbiol.* 6,  
516 245-252. <https://doi.org/10.1038/nrmicro1852>
- 517 Bruneel O., Pascault N., Egal M., Bancon-Montigny C., Goni-Urriza M.S., Elbaz-Poulichet F.,  
518 Personné J.-C., Duran R., 2008. Archaeal diversity in a Fe-As rich acid mine drainage at  
519 Carnoules (France). *Extremophiles* 12, 563-571.  
520 <https://doi.org/10.1007/s00792-008-0160-z>
- 521 Cabral L., Júnior G.V.L., Pereira De Sousa S.T., Franco Dias A.C., Cadete L.L., Dini  
522 Andreote F., Hess M., de Oliveira V.M., 2016. Anthropogenic impact on mangrove  
523 sediments triggers differential responses in the heavy metals and antibiotic resistomes of  
524 microbial communities. *Environ. Pollut.* 216, 460-469.  
525 <https://doi.org/10.1016/j.envpol.2016.05.078>
- 526 Chahal H.K., Dai Y., Saini A., Ayala-Castro C., Outten F.W., 2009. The *SufBCD* Fe-S  
527 scaffold complex interacts with *SufA* for Fe-S cluster transfer. *Biochemistry* 48,  
528 10644-10653. <https://doi.org/10.1021/bi901518y>
- 529 Chen Y.T., Li J.T., Chen L.X., Hua Z.S., Huang L.N., Liu J., Xu B.B., Liao B., Shu W.S.,  
530 2014. Biogeochemical processes governing natural pyrite oxidation and release of acid

531 metalliferous drainage. *Environ. Sci. Technol.* 48, 5537-5545.  
532 <https://doi.org/10.1021/es500154z>

533 Chen Z., Liu W., Zhong X., Fei Y., He H., Ding K., Chao Y., Tang Y., Wang S., Qiu R., 2021.  
534 Genome- and community-level interaction insights into the ecological role of archaea in  
535 rare earth element mine drainage in South China. *Water Res.* 201, 117331.  
536 <https://doi.org/10.1016/j.watres.2021.117331>

537 Cravo-Laureau C., Cagnon C., Lauga B., Duran R., 2017. *Microbial Ecotoxicology*. Springer  
538 International Publishing, Switzerland. <https://doi.org/10.3389/fmicb.2020.01342>

539 Douglas G.M., Maffei V.J., Zaneveld J., Yurgel S.N., Brown J.R., Taylor C.M., Huttenhower  
540 C., Langille M.G.I., 2019. PICRUSt2: an improved and extensible approach for  
541 metagenome inference. *bioRxiv* <http://dx.doi.org/10.1101/672295>

542 Frenk S., Hadar Y., Minz D., 2018. Quality of irrigation water affects soil functionality and  
543 bacterial community stability in response to heat disturbance. *Appl. Environ. Microbiol.*  
544 84, e02087-17. <https://doi.org/10.1128/AEM.02087-17>

545 Garg N. and Kaur H., 2013. Response of antioxidant enzymes, phytochelatins and glutathione  
546 production towards Cd and Zn stresses in *Cajanus cajan* (L.) Millsp. genotypes  
547 colonized by arbuscular mycorrhizal fungi. *J. Agron. Crop Sci.* 199, 118-133.  
548 <https://doi.org/10.1111/j.1439-037X.2012.00533.x>

549 Giloteaux L., Duran R., Casiot C., Bruneel O., Elbaz-Poulichet F., Goñi-Urriza M., 2013.  
550 Three-year survey of sulfate-reducing bacteria community structure in Carnoulès acid  
551 mine drainage (France), highly contaminated by arsenic. *FEMS Microbiol. Ecol.*  
552 383, 724-737. <https://doi.org/10.1111/1574-6941.12028>

553 Han Y., Wang J., Zhao Z, Chen J., Lu H., Liu G., 2017. Fishmeal application induces  
554 antibiotic resistance gene propagation in mariculture sediment. *Environ. Sci. Technol.* 51,  
555 10850-10860. <https://doi.org/10.1021/acs.est.7b02875>

556 Herren C.M. and McMahon K.D., 2018. Keystone taxa predict compositional change in  
557 microbial communities. *Environ Microbiol.* 20, 2207-2217.  
558 <https://doi.org/10.1111/1462-2920.14257>

559 Hudson-Edwards K., 2016. Tackling mine wastes. *Science* 352, 288-290.  
560 <http://doi.org/10.1126/science.aaf3354>

561 Hugoni M., Taib N., Debros D., Domaizon I., Dufournel I.J., Bronner G., Salter I., Agogué  
562 H., Mary I., Galand P.E., 2013. Structure of the rare archaeal biosphere and seasonal  
563 dynamics of active ecotypes in surface coastal waters. *PNAS* 110, 6004-6009.

564 <https://doi.org/10.1073/pnas.1216863110>

565 Jia T., Wang X., Guo T., Chai B., 2021. Litter decomposition of *Imperata cylindrica* in a  
566 copper tailing areas with different restoration history: Fungal community dynamics and  
567 driving factors. *Front. Microbiol.* 12, 780015.  
568 <https://doi.org/10.3389/fmicb.2021.780015>

569 Jia Y., Huang H., Chen Z., Zhu Y., 2014. Arsenic uptake by rice is influenced by  
570 microbe-mediated arsenic redox changes in the rhizosphere. *Environ. Sci. Technol.* 48,  
571 1001-1007. <https://doi.org/10.1021/es403877s>

572 Langille M.G.I., Zaneveld J., Caporaso J.G., McDonald D., Knights D., Reyes J.A., Clemente  
573 J.C., Burkepile D.E., Vega Thurber R.L., Knight R., Beiko R.G., Huttenhower C., 2013.  
574 Predictive functional profiling of microbial communities using 16S rRNA marker gene  
575 sequences. *Nat. Biotechnol.* 31, 814-821. <https://doi.org/10.1038/nbt.2676>

576 Latzel V., Allan E., Bortolini S.A., Colot V., Fischer M., Bossdorf O., 2013. Epigenetic  
577 diversity increases the productivity and stability of plant populations. *Nat. Commun.* 4,  
578 2875. <https://doi.org/10.1038/ncomms3875>

579 Lehtovirta-Morley L.E., Sayavedra-Soto L.A., Gallois N., Schouten S., Stein L.Y., Prosser  
580 J.I., Nicola G.W., 2016. Identifying potential mechanisms enabling acidophily in the  
581 ammonia-oxidizing archaeon "*Candidatus Nitrosotalea devanattera*." *Appl Environ*  
582 *Microbiol.* 82, 2608-2619. <https://doi.org/10.1128/aem.04031-15>

583 Liang J., Liu J., Jia P., Yang T., Zeng Q., Zhang S., Liao B., Shu W., Li J., 2020. Novel  
584 phosphate-solubilizing bacteria enhance soil phosphorus cycling following ecological  
585 restoration of land degraded by mining. *ISME J.* 14, 1600-1613.  
586 <https://doi.org/10.1038/s41396-020-0632-4>

587 Li M., Yao J., Sunahara G., Hawari J., Duran R., Liu J., Liu B., Cao Y., Pang W., Li H., Li Y.,  
588 Ruan Z., 2022. Novel microbial consortia facilitate metalliferous immobilization in  
589 non-ferrous metal(loid)s contaminated smelter soil: efficiency and mechanisms. *Environ*  
590 *Pollut.* 15, 120042. <https://doi.org/10.1016/j.envpol.2022.120042>

591 Liu B., Yao J., Chen Z., Ma B., Liu J., Li H., Zhu X., Li M., Cao Y., Pang W., Zhao C.,  
592 Mihucz V.G., Duran R., 2022a. Unraveling ecological risk of As/Sb and other  
593 metal(loid)s and fungal community responses in As/Sb smelting-intensive zone: A  
594 typical case study of Southwest China. *J. Clean. Prod.* 338, 130525.  
595 <https://doi.org/10.1016/j.jclepro.2022.130525>

596 Liu B., Yao J., Ma B., Li S., Duran R., 2022b. Disentangling biogeographic and underlying  
597 assembly patterns of fungal communities in metalliferous mining and smelting soils. *Sci.*



- 598 *Total Environ.* 845, 157151. <https://doi.org/10.1016/j.scitotenv.2022.157151>
- 599 Liu J., Yao J., Duran R., Mihucz V.G., Hudson-Edwards K.A., 2019a. Bacterial shifts during  
600 in-situ mineralization bio-treatment to non-ferrous metal(loid) tailings. *Environ. Pollut.*  
601 255, 113165. <https://doi.org/10.1016/j.envpol.2019.113165>
- 602 Liu J., Yao J., Wang F., Ni W., Liu X., Sunahara G., Duran R., Jordan G., Hudson-Edwards  
603 K.A., Alakangas L., Solevic-Knudsen T., Zhu X., Zhang Y., Li Z., 2018. China's most  
604 typical nonferrous organic-metal facilities own specific microbial communities. *Sci.*  
605 *Rep-UK* 8, 12570-12580. <https://doi.org/10.1038/s41598-018-30519-1>
- 606 Liu J., Yao J., Wang F., Min N., Gu J., Li Z., Sunahara G., Duran R., Solevic-Knudsen T.,  
607 Hudson-Edwards K.A., Alakangas L., 2019b. Bacterial diversity in typical abandoned  
608 multi-contaminated nonferrous metal(loid) tailings during natural attenuation. *Environ.*  
609 *Pollut.* 247, 98-107. <https://doi.org/10.1016/j.envpol.2018.12.045>
- 610 Liu J., Yao J., Zhou D., Liu B., Liu H., Li M., Zhao C., Sunahara G., Duran R., 2023.  
611 Mining-related multi-resistance genes in sulfate-reducing bacteria treatment of typical  
612 karst nonferrous metal(loid) mine tailings in China. *Environ. Sci. Pollut. R.* 30,  
613 104753-104766. <https://doi.org/10.1007/s11356-023-29203-3>
- 614 Liu J., Yao J., Zhu X., Zhou D., Duran R., Mihucz V.G., Bashir S., Hudson-Edwards K.A.,  
615 2021. Metagenomic exploration of multi-resistance genes linked to microbial attributes  
616 in active nonferrous metal(loid) tailings. *Environ. Pollut.* 273, 115667.  
617 <https://doi.org/10.1016/j.envpol.2020.115667>
- 618 Liu S., García-Palacios P., Tedersoo L., Guirado E., van der Heijden M.G.A., Wagg C., Chen  
619 D., Wang Q., Wang J., Singh B.K., Delgado-Baquerizo M., 2022c. Phylotype diversity  
620 within soil fungal functional groups drives ecosystem stability. *Nat. Ecol. Evol.* 6,  
621 900-909. <https://doi.org/10.1038/s41559-022-01756-5>
- 622 **Lloyd C., 2023. Cultured microbial diversity of Parys mountain acidic site. PhD thesis.**  
623 **School of Environmental and Natural Sciences, UK: Bangor University.**
- 624 Luo C., Routh J., Luo D., Wei L., Liu Y., 2021. Arsenic in the Pearl River Delta and its  
625 related waterbody, South China: occurrence and sources, a review. *Geosci. Lett.* 8, 12.  
626 <https://doi.org/10.1186/s40562-021-00185-9>
- 627 Malard L.A., Mod H.K., Guex N., Broennimann O., Yashiro E., Lara E., Mitchell E.A.,  
628 Niculita-Hirzel H., Guisan A., 2021. Comparative analysis of diversity and  
629 environmental niches of soil bacterial, archaeal, fungal and protist communities reveal  
630 niche divergences along environmental gradients in the Alps. *Soil Biol. Biochem.* 169,  
631 108674. <https://doi.org/10.1016/j.soilbio.2022.108674>

632 Mamet S.D., Lamb E.G., Piper C.L., Winsley T., Siciliano S.D., 2017. Archaea and bacteria  
633 mediate the effects of native species root loss on fungi during plant invasion. *ISME J.* 11,  
634 1261-1275. <https://doi.org/10.1038/ismej.2016.205>

635 Maynaud G., Brunel B., Yashiro E., Mergeay M., Cleyet-Marel J.C., Le Quéré A., 2014.  
636 *CadA* of *Mesorhizobium metallidurans* isolated from a zinc-rich mining soil is a  
637 P(IB-2)-type ATPase involved in cadmium and zinc resistance. *Res Microbiol.* 165,  
638 175-89. <https://doi.org/10.1016/j.resmic.2014.02.001>

639 Mendez-Garcia C., Pelaez A. I., Mesa V., Sánchez J., Golyshina O.V., Ferrer M., 2015.  
640 Microbial diversity and metabolic networks in acid mine drainage habitats. *Front.*  
641 *Microbiol.* 6, 475. <https://doi.org/10.3389/fmicb.2015.00475>

642 Moussa T.A.A., Al-Zahrani H.S., Almaghrabi O.A., Abdelmoneim T.S., Fuller M.P., 2017.  
643 Comparative metagenomics approaches to characterize the soil fungal communities of  
644 western coastal region, Saudi Arabia. *PLOS ONE* 12, e0185096.  
645 <https://doi.org/10.1371/journal.pone.0185096>

646 Nguyen N.H., Song Z., Bates S.T., Branco S., Tedersoo L., Menke J., Schilling J.S., Kennedy  
647 P.G., 2016. FUNGuild: an open annotation tool for parsing fungal community datasets  
648 by ecological guild. *Fungal Ecol.* 20, 241-248.  
649 <https://doi.org/10.1016/j.funeco.2015.06.006>

650 Ning D., Yuan M., Wu L., Zhang Y., Guo X., Zhou X., Yang Y., Arkin A.P., Firestone M.K.,  
651 Zhou J., 2020. A quantitative framework reveals ecological drivers of grassland  
652 microbial community assembly in response to warming. *Nat. Commun.* 11, 4717.  
653 <https://doi.org/10.1038/s41467-020-18560-z>

654 Rahman M.S., Hossain M.S., Saha S.K., Rahman S., Sonne C., Kim K., 2021. Homology  
655 modeling and probable active site cavity prediction of uncharacterized arsenate reductase  
656 in bacterial spp. *Appl. Biochem. Biotechnol.* 193, 3119-3135.  
657 <https://doi.org/10.1007/s12010-020-03392-w>

658 Ren S., Li Q., Xie L., Xie J., 2017. Molecular mechanisms underlying the function diversity  
659 of *ArsR* family metalloregulator. *Crit. Rev. Eukar. Gene.* 27, 19-35.  
660 <https://doi.org/10.1615/critreveukaryotgeneexpr.2016018476>

661 Samaddar S., Karp D. S., Schmidt R., Devarajan N., McGarvey J.A., Pires A.F.A., Scow K.,  
662 2021. Role of soil in the regulation of human and plant pathogens: soils' contributions to  
663 people. *Phil. Trans. R. Soc. B* 376, 20200179. <https://doi.org/10.1098/rstb.2020.0179>

664 Sheridan P.O., Meng Y., Williams T.A., Gubry-Rangin C., 2022. Recovery of  
665 *Lutacidiplasmatales* archaeal order genomes suggests convergent evolution in

666 *Thermoplasmatota*. *Nat. Commun* 13, 4110. <https://doi.org/10.1038/s41467-022-31847-7>

667 Sloan W.T., Lunn M., Woodcock S., Head I.M., Nee S., Curtis T.P., 2006. Quantifying the  
668 roles of immigration and chance in shaping prokaryote community structure. *Environ.*  
669 *Microbiol.* 8, 732-740. <https://doi.org/10.1111/j.1462-2920.2005.00956.x>

670 Speer R.M., Zhou X., Volk L.B., Liu K.J., Hudson L.G., 2023. Arsenic and cancer: Evidence  
671 and mechanisms. *Adv. Pharmacol.* 96, 151-202.  
672 <https://doi.org/10.1016/bs.apha.2022.08.001>

673 Sun W., Xiao E., Haggblom M., Krumins V., Dong Y., Sun X., Li F., Wang Q., Li B., Yan B.,  
674 2018. Bacterial survival strategies in an alkaline tailing site and the physiological  
675 mechanisms of dominant phylotypes as revealed by metagenomic analyses. *Environ. Sci.*  
676 *Technol.* 52, 13370-13380. <https://doi.org/10.1021/acs.est.8b03853>

677 Sunagawa S., Coelho L. P., Chaffron S. *et al.*, 2015. Structure and function of the global  
678 ocean microbiome. *Science* 348, 6237. <https://hal.science/hal-01233742>

679 Suting E.G. and Olivia Devi N., 2021. Occurrence and diversity of arbuscular mycorrhizal  
680 fungi in trap cultures from limestone mining sites and un-mined forest soil of Mawsmai,  
681 Meghalaya. *Trop. Ecol.* 62, 525-537. <https://doi.org/10.1007/s42965-021-00144-7>

682 Taleski V., Dimkić I., Boev B., Boev I., Živković S., Stanković S., 2020. Bacterial and fungal  
683 diversity in the lorandite (TlAsS<sub>2</sub>) mine 'Allchar' in the Republic of North Macedonia.  
684 *FEMS Microbiol. Ecol.* 96, 155. <https://doi.org/10.1093/femsec/fiaa155>

685 Torres S.B., Petrik A., Szabo K.Z., Jordan G., Yao J., Szabó C., 2018. Spatial relationship  
686 between the field-measured ambient gamma dose equivalent rate and geological  
687 conditions in a granitic area, Velence Hills, Hungary: an application of digital spatial  
688 analysis methods. *J. Environ. Radioactiv.* 192, 267-278.  
689 <https://doi.org/10.1016/j.jenvrad.2018.07.001>

690 Tuhý M., Hrstka T., Ettler V., 2020. Automated mineralogy for quantification and  
691 partitioning of metal(loid)s in particulates from mining/smelting-polluted soils. *Environ.*  
692 *Pollut.* 266, 115118. <https://doi.org/10.1016/j.envpol.2020.115118>

693 Umezawa K., Kojima H., Kato Y., Fukui M., 2020. Disproportionation of inorganic sulfur  
694 compounds by a novel autotrophic bacterium belonging to *Nitrospirota*. *Syst. Appl.*  
695 *Microbiol.* 43, 126110. <https://doi.org/10.1016/j.syapm.2020.126110>

696 Vasconcelos A.L.S., Defalco T., Dias A.C.F., Barrientos L., Bernardino A.F., Andreote F.D.,  
697 Núñez-Montero K., 2021. Complete genome sequence of a *Mucilaginibacter* sp. strain  
698 isolated from estuarine soil contaminated with mine tailings from the Samarco Disaster

699 at Fundão Dam. *Microbiol. Resour. Announc.* 10, e0077921.  
700 <https://doi.org/10.1128/mra.00779-21>

701 Wang L., Ji B., Hu Y., Liu R., Sun W., 2017. A review on in situ phytoremediation of mine  
702 tailings. *Chemosphere* 184, 594-600. <https://doi.org/10.1016/j.chemosphere.2017.06.025>

703 Wang Y., Dong R., Zhou Y., Luo X., 2019. Characteristics of groundwater discharge to river  
704 and related heavy metal transportation in a mountain mining area of Dabaoshan,  
705 Southern China. *Sci. Total Environ.* 679, 346-358.  
706 <https://doi.org/10.1016/j.scitotenv.2019.04.273>

707 Xiao E., Krumins V., Dong Y., Xiao T., Ning Z., Xiao Q., Sun W., 2016. Microbial diversity  
708 and community structure in an antimony-rich tailings dump. *Appl. Microbiol. Biotechnol.*  
709 100, 7751-7763. <https://doi.org/10.1007/s00253-016-7598-1>

710 Yi X.Y., Yang Y.P., Yuan H.Y., Chen Z., Suan G., Zhu Y., 2019. Coupling metabolisms of  
711 arsenic and iron with humic substances through microorganisms in paddy soil. *J. Hazard.*  
712 *Mater.* 373, 591-599. <https://doi.org/10.1016/j.jhazmat.2019.03.113>

713 Yi X., Liang J., Su J., Jia P., Lu J., Zheng J., Wang Z., Feng S., Luo Z., Ai H., Liao B., Shu  
714 W., Li J., Zhu Y., 2022. Globally distributed mining-impacted environments are  
715 underexplored hotspots of multidrug resistance genes. *ISME J.* 16, 2099-2113.  
716 <https://doi.org/10.1038/s41396-022-01258-z>

717 Zhang Y., Wang F., Hudson-Edwards K., Blake R., Zhao F., Yuan Z., Gao W., 2020.  
718 Characterization of mining-related aromatic contaminants in active and abandoned  
719 metal(loid) tailings ponds. *Environ. Sci. Technol.* 54, 15097-15107.  
720 <https://doi.org/10.1021/acs.est.0c03368>

721 Zhu X., Yao J., Wang F., Yuan Z., Liu J., Jordan G., Šolević Knudsen T., Avdalović J., 2018.  
722 Combined effects of antimony and sodium diethyldithiocarbamate on soil microbial  
723 activity and speciation change of heavy metals. Implications for contaminated lands  
724 hazardous material pollution in nonferrous metal mining areas. *J. Hazard. Mater.* 349,  
725 160-167. <https://doi.org/10.1016/j.jhazmat.2018.01.044>

726 Zhu J., Yang K., Chen Y., Fan G., Zhang L., 2021. Revealing the substitution preference of  
727 zinc in ordinary Portland cement clinker phases: a study from experiments and DFT  
728 calculations. *J. Hazard. Mater.* 409, 124504.  
729 <https://doi.org/10.1016/j.jhazmat.2020.124504>

730 **Figure legends**

731 **Fig. 1** SEM images of collected tailings samples, and XRD analysis of collected  
732 tailings samples. a: Quartz (PDF# 46-1045), b: Manganese copper oxide  
733 (PDF# 82-0668), c: Iwakiite (PDF# 38-0430), d: Celadonite (PDF# 17-0521),  
734 e: Magnesium iron oxide (PDF# 77-2367).

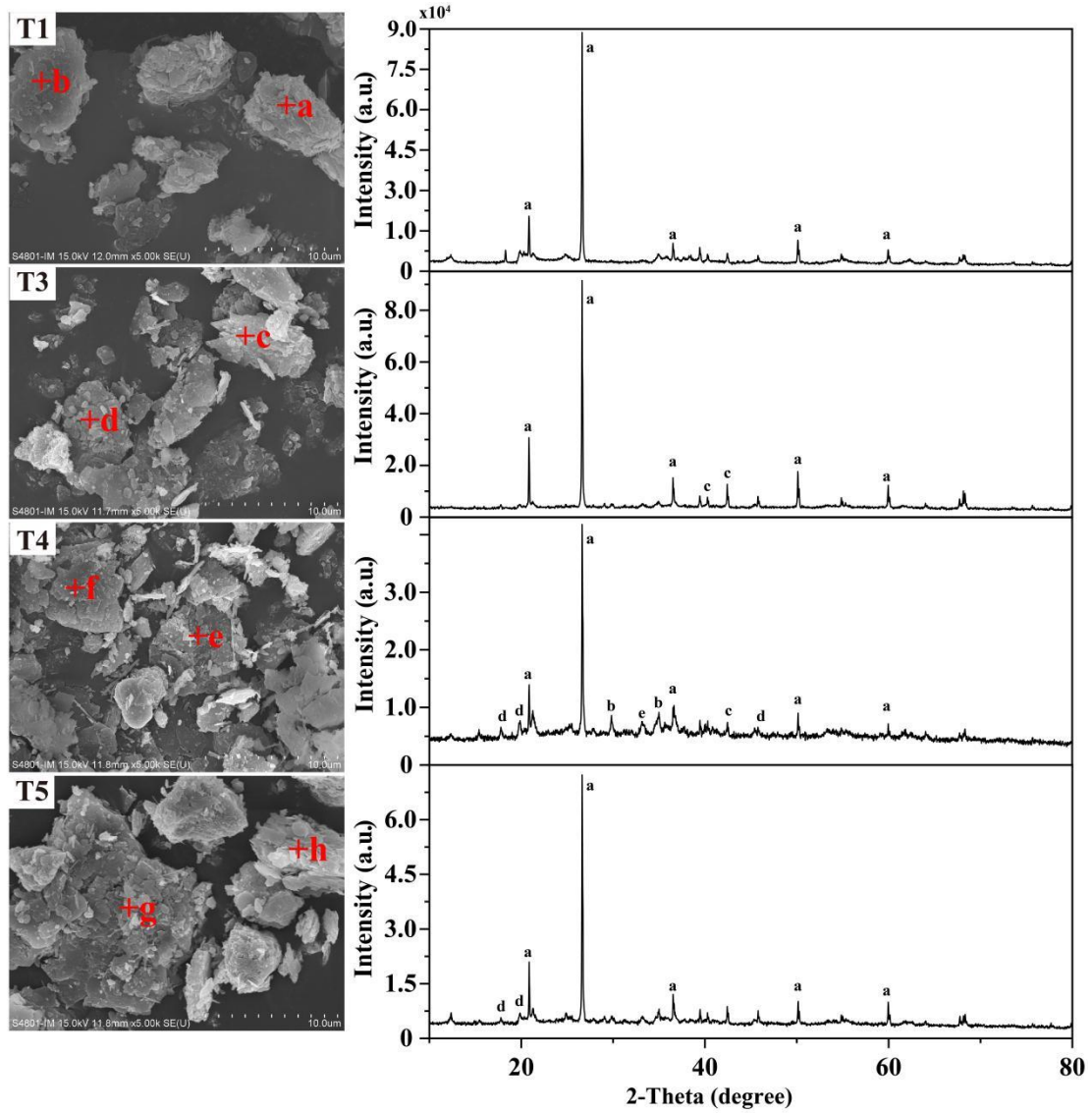
735 **Fig. 2** General patterns of microbial alpha-diversity and community assembly. (a)  
736 Alpha-diversity of bacterial, fungal, and archaeal in the sampling sites.  
737 Significant difference (based on the Wilcoxon test;  $p \leq 0.05$ ). Relative  
738 abundance of (b) bacterial, (c) fungal, and (d) archaeal communities. Bubble  
739 sizes represent the relative abundance of the genera. The bubble color is  
740 colored according to the classification information.

741 **Fig. 3** General patterns of microbial beta-diversity in studied tailings. (a-d)  
742 Respective beta-diversity of bacterial, fungal, archaeal, and all taxa  
743 communities based on non-metric multidimensional scaling (NMDS)  
744 analysis. (e) Partial Mantel test for co-relation of three microbial groups and  
745 physical-chemical parameters. A color gradient denotes the Pearson's  
746 correlation coefficients.

747 **Fig. 4** Metacommunity network analysis for the detected bacterial, fungal, and  
748 archaeal taxa in tailings. Each node indicates one OTU. The corresponding  
749 species of each OTU was listed in Table S6. The colors of the node represent  
750 the different major phyla and genera. The size of the species-node denotes  
751 the abundance of species. Nodes unnamed represented the  
752 no\_rank/un-classified genera. The red line represents the positive correlation  
753 between the two individual nodes, and the green line shows a negative  
754 (exclusion) correlation.

755 **Fig. 5** (a) Variation of incidence-based (Raup-Crick) beta-diversity for microbial  
756 communities. (b) Comparison of the maximum likelihood fit of the neutral  
757 models estimated by the Akaike information criterion for microbial  
758 communities in tailings. (c) Comparison of mean habitat niche breadths in all  
759 taxa in bacterial, fungal, and archaeal communities.

760 **Fig. 6** Independent-samples *T*-test and correlation of KOs with geochemical  
761 parameters. (a) Difference analysis for KOs related to metals of bacterial and  
762 archaeal communities based on the independent-sample *T*-test. Data for the  
763 23 shared KOs were used for analysis. \* $0.01 < p \leq 0.05$ , \*\* $0.001 < p \leq 0.01$ ,  
764 \*\*\*  $p \leq 0.001$ . (b and c) Redundancy analyses (RDA) of geochemical  
765 parameters (arrows) and KEGG pathways (\*) of bacterial and archaeal  
766 communities, respectively.



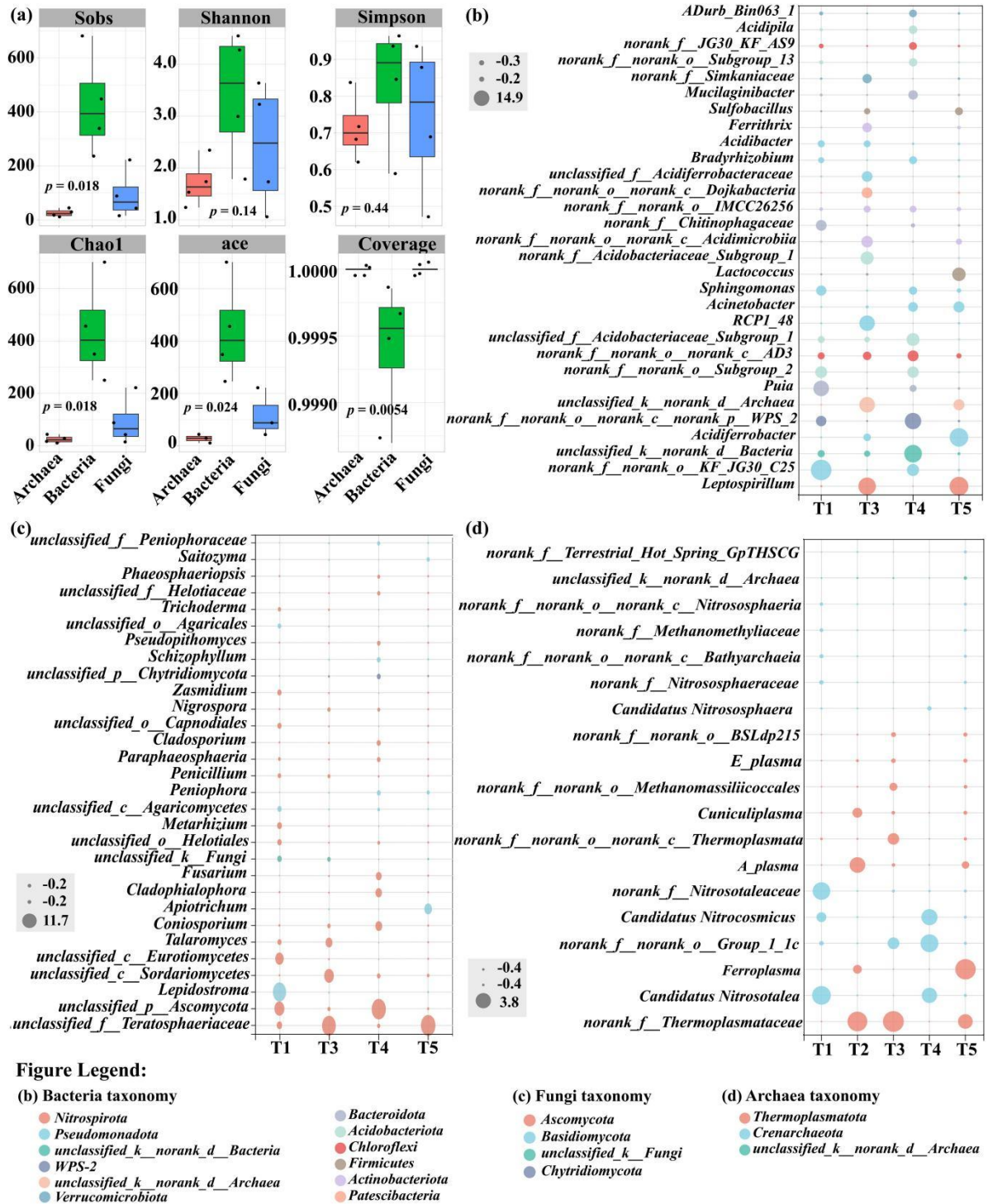
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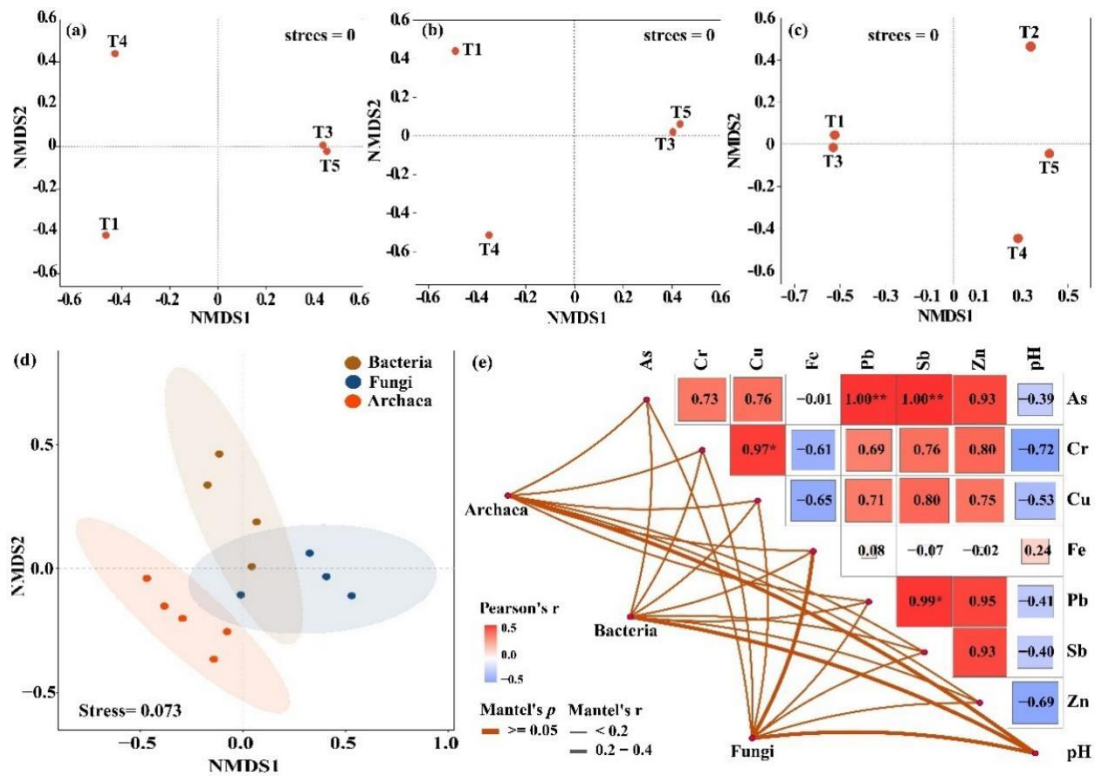
770 (PDF# 82-0668), c: Iwakiite (PDF# 38-0430), d: Celadonite (PDF# 17-0521),

771 e: Magnesium iron oxide (PDF# 77-2367).



772 **Fig. 2** General patterns of microbial alpha-diversity and community assembly. (a)  
 773 Alpha-diversity of bacterial, fungal, and archaeal in the sampling sites.  
 774 Significant difference:  $p \leq 0.05$  (analysis was based on the Wilcoxon test).  
 775 (b-d) Relative abundance of microorganisms, corresponding to bacterial,  
 776 fungal, and archaeal communities, respectively. Bubble size represented the  
 777 relative abundance of the genus. The bubble color indicates the classification  
 778 information.





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780 **Fig. 3** General patterns of microbial beta-diversity in studied tailings. (a-d)

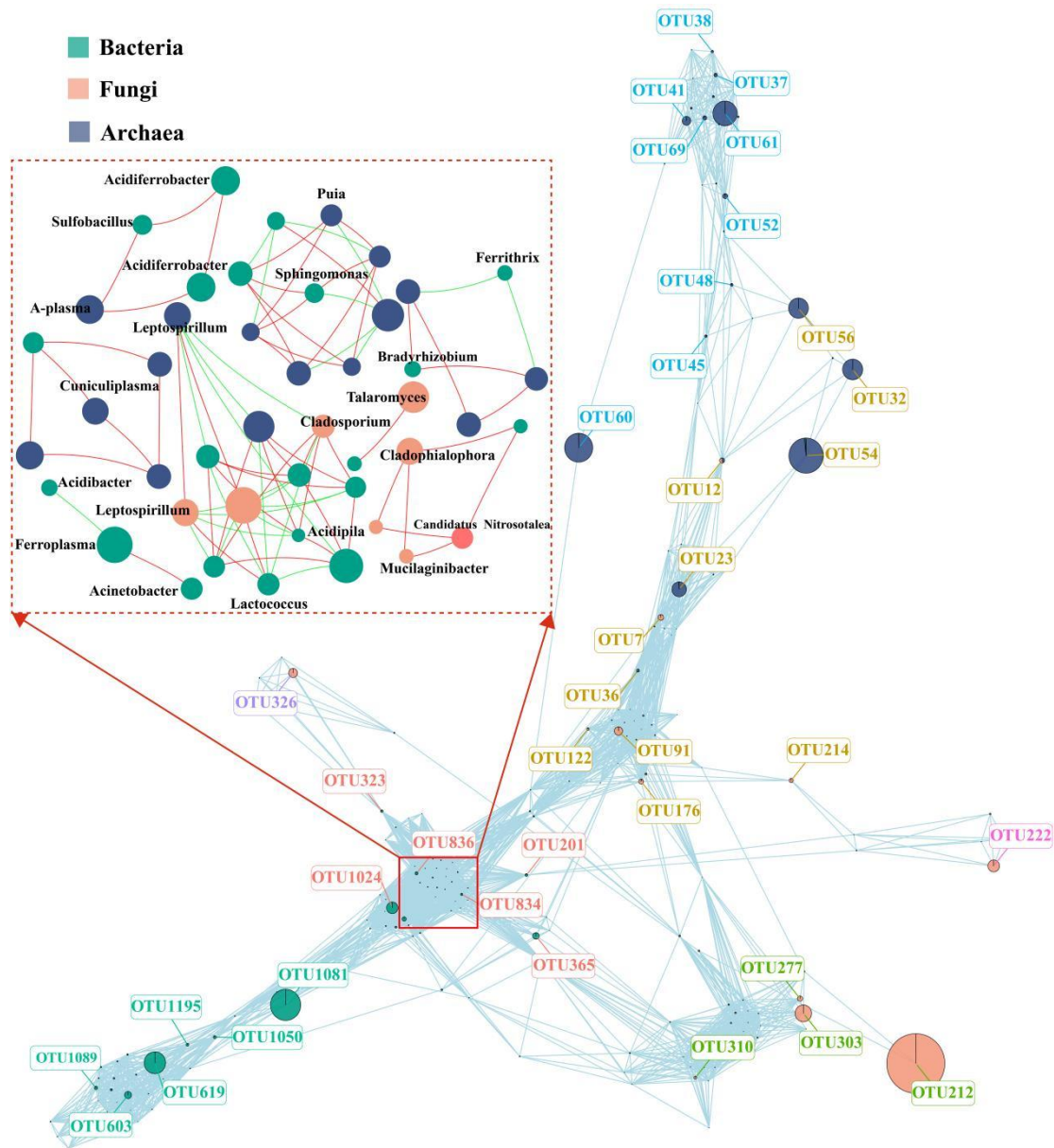
781 Beta-diversity of bacterial, fungal, archaeal, and all taxa communities based

782 on non-metric multidimensional scaling (NMSDS) analysis. (e) Partial Mantel

783 test for correlation of three microbial groups and physical-chemical

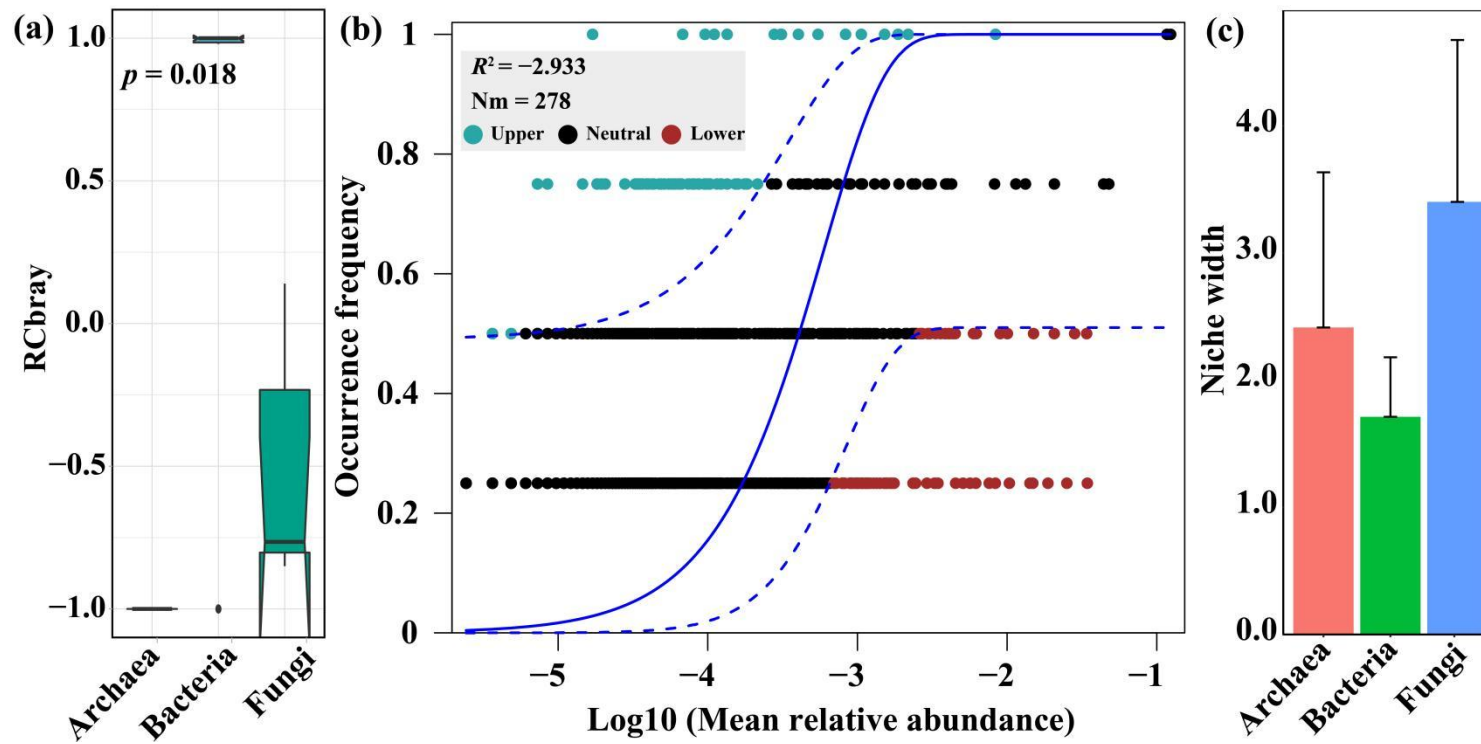
784 parameters. A color gradient denotes the Pearson's correlation coefficients.

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787 **Fig. 4** Metacommunity network analysis for the detected bacterial, fungal, and  
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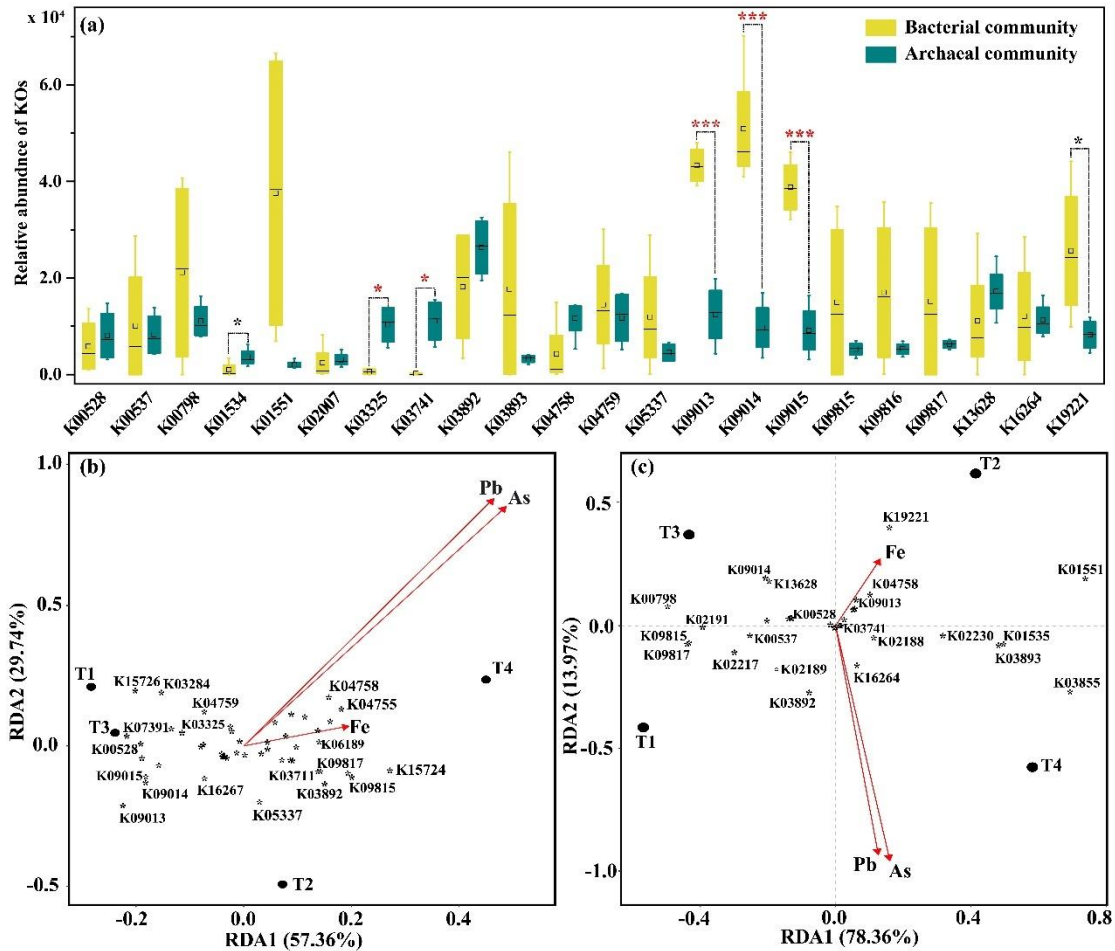
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