

Biological, geological and chemical effects of oxygen injection in underground gas storage aquifers in the setting of biomethane deployment

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1	Biological, Geological and Chemical Effects of Oxygen Injection
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3	Biomethane Deployment
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25 Abstract

The last few years have seen the proliferation of anaerobic digestion plants to produce 26 biomethane. Oxygen (O₂) traces added to biogas during the desulfurization process are co-27 injected in the gas network and can be stored in Underground Gas Storage (UGS). However, 28 there are no data available for the undesirable effects of O₂ on these anoxic environments, 29 especially on deep aquifers. In addition to mineral alteration, O₂ can have an impact on the 30 anaerobic autochthonous microbial life. In our study, the storage conditions of an UGS 31 aquifer were reproduced in a high-pressure reactor and bio-geo-chemical interactions between 32 the aqueous, gas and solid phases were studied. Sulfate was depleted from the liquid phase for 33 34 three consecutive times during the first 130 days of incubation reproducing the storage conditions (36 °C, 60 bar, methane with 1% CO₂). Sulfate-reducers, 35 such as Desulfovibrionaceae, were identified from the high-pressure system. Simulations with 36 37 PHREEQC were used to determine the thermodynamic equilibrium to confirm any gas consumption. CO₂ quantities decreased in the gas phase, suggesting its use as carbon source 38 by microbial life. Benzene and toluene, hydrocarbons found in traces and known to be 39 biodegradable in storages, were monitored and a decrease of toluene was revealed and 40 associated to the Peptococcaceae family. Afterwards, O2 was added as 1% of the gas phase, 41 42 corresponding to the maximum quantity found in biomethane after desulfurization process. Re-oxidation of sulfide to sulfate was observed along with the end of sulfate reducing activity 43 and toluene biodegradation and the disappearance of most of the community. H₂ surprisingly 44 appeared and accumulated as soon as hydrogenotrophic sulfate-reducers decreased. H₂ would 45 be produced via the necromass fermentation accomplished by microorganisms able to resist 46 the oxic conditions of $4.42 \cdot 10^{-4}$ mol.Kgw⁻¹ of O₂. The solid phase composed essentially of 47 quartz, presented no remarkable changes. 48

- 50 Keywords: deep aquifers, biomethane, oxygen, deep subsurface, high pressure, UGS,
- 51 geological storage

53 1. Introduction

Oxygen (O₂) can be injected in deep underground for many purposes. Its ability to promote 54 aerobic microbial metabolism is a key factor to achieve the endpoint in multiple current 55 56 practices especially during the application of enhanced oil recovery techniques on oil reservoir (Cai et al., 2015; Liu et al., 2018; Pannekens et al., 2019) and the bio-remediation of 57 groundwater (EPA, 2017; Lien et al., 2016; Trulli et al., 2016). However, despite its 58 59 aforementioned beneficial applications, the presence of oxygen can be deleterious during other procedures such as Carbon Capture and Storage (CCS) (Morgan et al., 2017; Wei et al., 60 2015), compressed air energy storage (CAES) (Guo et al., 2021; Kushnir et al., 2012) and 61 62 underground natural gas storage (UGS). Many studies have been undergone in the context of CCS storage in order to define the effect of carbon dioxide, but also the effect of oxygen 63 64 injected as an impurity on the geological formations. Similarly, the deep underground storage of biomethane can also involve co-injection of O₂. Indeed, O₂ is added to the biomethane 65 during sulfide reoxidation at the outlet of the anaerobic digestion plant. Depending on the 66 67 initial sulfide quantity, the proportion of O₂ may attain 1.5 % of the final gas mixture (Barik et 68 al., 2013). For underground gas storage, France sets a maximal O_2 proportion of 100 ppm, whereas Germany sets a maximal of 10 ppm. 69

70 Biomethane is an attractive alternative to fossil fuels particularly because biogas produced from biomass, such as green waste and sludge from wastewater, prevents re-circulating 71 carbon from oil, natural gas and coal. With a reported production of 1.2 billion m³ in 2015, 72 Europe is the world's leading biomethane producer. The lead application of the produced 73 biomethane consists of its use as vehicle fuel or its injection into the natural gas grid (Scarlat 74 75 et al., 2018). With the fluctuating and unpredictable demand throughout the year, biomethane storage is recently gaining an increasing interest. Due to their large storage capacity and their 76 ability to preserve gas under pressure, deep aquifers emerge as attractive underground storage 77

systems. Deep Underground Gas Storages (UGS) such as deep aquifers, are known for being 78 79 anoxic environments that form large ecosystems and harbor a much-diverse anaerobic microbial life (Magnabosco et al., 2018). The biodegradation of hydrocarbons was notably 80 studied in this setting, especially the anaerobic degradation of BTEX (Benzene, Toluene, 81 Ethylbenzene and Xylenes isomers) by autochthonous microorganisms in these ecosystems 82 (Aüllo et al., 2016; Bombach et al., 2010; Ranchou-Peyruse et al., 2017). Sulfate-reducing 83 84 bacteria (SRB) are particularly known to be active and to be key players in such anoxic environments (Berlendis et al., 2010; Bombach et al., 2010; Detmers et al., 2004; Fry et al., 85 1997; Itävaara et al., 2011; Ranchou-Peyruse et al., 2019). Part of the O₂ co-injected with the 86 87 biomethane during its storage can solubilize into the formation water leading to alterations of the preexisting thermodynamic equilibrium, which can result in drastic modifications in the 88 microbial communities and depletion of strictly anoxic micro-organisms. O₂ may also cause 89 90 mineral oxidation of the reservoir rock (Guo et al., 2021; Jung et al., 2013; Kushnir et al., 2012; Lu et al., 2016; Vu et al., 2018). For instance, the oxidation of pyrite (FeS₂) was 91 92 observed during CAES in different studies (Guo et al., 2021; Kushnir et al., 2012) and during 93 CCS (Jung et al., 2013; Shao et al., 2014). This alteration can cause a variation in the reservoir's porosity and permeability especially if large amounts of redox-sensitive minerals 94 were present (Lu et al., 2016). Dissolution of rocks, mineral precipitations and corrosion of 95 well materials were shown as possible (Zettlitzer et al., 2010; Wang et al., 2011). Variation in 96 the aqueous phase composition resulting from mineral dissolution can impact the 97 thermodynamic equilibrium. Identifying this equilibrium is important to determine the 98 99 material balance and to estimate any production or consumption of gas such as carbon dioxide consumption by the microbial life as a carbon source or oxygen consumption due to chemical 100 reactions. However, experimental solubility data corresponding to our aqueous and gas 101 compositions under aquifer's conditions are very limited. The studied conditions of the 102

system were a temperature of 36 °C and a total pressure of 60 bar, similar to the aquifer's real 103 104 storage conditions. Even though many CCS studies targeted storage effects on the site, multiple uncertainties remain regarding the effects of impurities namely O₂ during 105 106 biomethane storage in natural gas reservoirs such as aquifers. In order to elucidate more answers on this topic, a multi-disciplinary study is undergone and, to our knowledge, this is 107 one of few studies involving the reproduction of the aquifer's *in-situ* conditions in a reactor 108 109 using solid, liquid and microbial underground samples. During this study, the aqueous and gas compositions were followed up regularly via chromatography techniques. Samples from the 110 aqueous phase were recovered under anoxic conditions for microbial studies. Afterwards, O2 111 112 was injected at 1% of the gas mixture, corresponding to the maximum quantity found in biomethane. Before and in the end of the experiment, the solid phase composed of a reservoir 113 rock sample undergone mineral analyses. The subsequent bio-geo-chemical interactions 114 115 between the gas mixture, reservoir rock, formation water and harbored microbial life were determined under temperature and pressure of a selected aquifer UGS. 116

117

118 2. Materials and methods

119 2.1- Site description and formation water characteristics.

The studied reservoir is a UGS in the South Aquitaine sedimentary basin (582 m depth), in Southwest France used for natural gas storage since 1981. The well that was selected for this study is the closest to the gas bubble (about 250 m) and has already been identified in previous studies with the code name Ab_L_1 (Ranchou-Peyruse et al., 2019). Thus, sampled water is the most exposed for gases interactions due to its proximity to the gas bubble. Some of the physicochemical characteristics of the formation water (LCE-SOBEGI Laboratory, Lacq, France) and its overall composition were summarized in Table 1. Formation water

(0.634 L) was recovered from the well under pressure (January 2019) with a downhole 127 128 sampler, equipped with a PDS Sampler (Leutert Bottom Hole Positive Displacement Sampler) (Kampman et al., 2014). Wellhead water samples were also recovered. Samples were stored at 129 4°C until use. Wellhead samples were filter-sterilized at 0.1 µm (PES 47 mm membranes, 130 0.1 µm, Sartorius) while maintaining anoxic conditions: 0.854 L of these samples was added 131 to the formation water (0.634 L) to increase the volume, a month before the injection into the 132 high-pressure reactor. A volume of 0.1 L was taken from the final mixture for a microbial 133 diversity study, whereas the rest (1.39 L) was injected into the reactor. 134

The reservoir rock was composed of infra-molassic sands, with a porosity varying between 25 and 35 % (Ranchou-Peyruse et al., 2019). The rock samples were composed of cuttings collected by TEREGA from a drilling carried out near the studied site. X-ray diffraction analysis applied on the solid sample before the experiment showed that the solid phase was mainly composed of quartz with minor calcite content (Figure S1). Iron sulfides (pyrite and marcasite), iron oxi-hydroxides (maghemite and hematite) and clay minerals were also identified as trace minerals.

142 2.2- *High-pressure reactor*.

The apparatus was composed of an experimental cell presented in Figure 1. The autoclave 143 was made of Hastelloy C-276 in order to prevent any corrosion problems. A basket held the 144 solid phase and was composed of a metallic disc with pores of less than 10 µm in order to 145 prevent particles from falling to the bottom of the reactor without interfering with the 146 movement of microorganisms. The operating pressure and temperature can go up to 150 bar 147 and 150 °C respectively. A safety valve of 150 bar connected to a pressure compensation 148 chamber protected the overall system. An insulating coating with heating resistors heated the 149 autoclave. Two thermocouples permitted monitoring of the liquid and gas phase temperatures 150 with a precision of ± 1 °C. A double jacket was used to maintain the temperature of the 151

experimental cell. The cell pressure was monitored by the Keller PA(A)-33X pressure gauge 152 with a precision of 0.3 bar. In order to maintain a constant pressure throughout the 153 experiment, a mobile piston was placed into the reactor. A 1L capacity pump allowed the 154 piston to move upwards in order to increase the cell pressure and downwards to decrease it. 155 The pump can operate at a maximum pressure of 150 bar and was monitored by a pressure 156 gauge with a precision of 0.05 %. A flow rate of up to 220 mL/min can be imposed with a 157 158 precision of 0.1 % of the set point. The overall autoclave volume was 5.41 L with an available volume of 4.2 L when the piston was at its lowest level. The solid basket had a volume of 0.8 159 L. The system was mixed using a stirring shaft operated by a variable speed motor with a 160 161 maximum stirring speed of 500 rpm. A Rushton turbine was used for the liquid phase and a double disc stirrer with four vertical blades for the gas phase. The composition of the gas 162 mixture injected at the beginning of the biotic experiment was methane, 1 % of carbon 163 164 dioxide, 7.95 ppm of benzene and 3.57 ppm of toluene. The oxygen added to simulate biomethane injection had a purity of >99.995% (CAS: 7782-55-7). 165

The experiment ran over 180 days. Initially, a volume of 1.39 L of aqueous phase was injected in the reactor and $4.57 \pm 2.2 \cdot 10^{-1}$ mol of gas mixture. The system liquid-gas-solidbiomass was under development for 130 days under a total pressure of 60 bar and a temperature of 36 °C. The oxygen injection (1 % of 60 bar) occurred at day 131. The Figure 2 presents the analysis undergone through the experiment at different timings.

171 2.3- Abiotic experiment

A preliminary abiotic experiment was initially undergone in order to detect the liquid-solid interactions before adding the microbial life to the system. 500 mL of distilled water were mixed with 100 mL of sand reservoir in a sealed 1L bottle that was subsequently flushed with N₂ gas at atmospheric pressure. The composition of the aqueous phase was followed in order to detect compounds dissolution from the solid into the liquid phase.

177 2.4- Analytical analyses.

The liquid phase was analyzed using an ionic chromatography with a variable wavelength 178 detector from ThermoFisher Scientific (Dionex Integrion HPIC). The column IonPac AS28-179 180 Fast-4µm (4x150 mm) was used with the AG28-4µm (4x30 mm) guard-column for anion detection, particularly fluoride, chloride, sulfate, acetate, carbonate/bicarbonate and sulfide. 181 The cations were quantified with the Dionex Easion System (ThermoFisher Scientific). The 182 183 column IC Dionex IonPac CS12A (4x250 mm) was used with the guard-column CG12A (4x50 mm) for sodium, potassium, magnesium and calcium. A liquid sample was taken from 184 the liquid output using a syringe and stored in a flask that was sealed with rubber. Air was 185 immediately extracted from the flask using N₂ gas flush. The sample (1 mL) was directly 186 injected into the instrument. The analyses were performed in triplicate. The supposed 187 measurement uncertainty on both techniques was $\pm 5\%$. 188

The gas phase composition was monitored by a gas chromatography in line with a micro 189 190 thermal conductivity detector (GC-µTCD; Micro GC Fusion; Chemlys; France). The 191 'Molsieve 5A' column was used with helium as a carrier gas for quantification of methane, hydrogen, oxygen and nitrogen. The column 'Plot Q' was used with argon as carrier gas for 192 quantification of carbon dioxide and hydrogen sulfide. The instrument was connected to the 193 194 reactor's valve. Once open, the sample traversed a pressure compensation tube to reach the piston where it attained the atmospheric pressure and it undergone analysis at that moment. 195 196 All measures were obtained after three analyses, and the uncertainty was ± 2 %. The obtained 197 results were used to calculate the quantity of each component in the gas phase. The non-ideal gas equation was used to calculate the total number of moles in the autoclave at the moment 198 199 of analysis. The selected compressibility factor was 0.9 (calculated with PHREEQC using 200 phreeqc.dat as database).

The identification and quantification of benzene and toluene in the liquid and gas phases were 201 202 performed using a ThermoFisher Scientific gas chromatograph coupled with a quadrupole spectrometer (GC-MS) (ISQ QD Single Quadrupole MS 203 - Trace 1310). mass Preconcentration of benzene toluene performed by 204 and was SPME with a polydimethylsiloxane/carboxene (PDMS/CAR) fiber and the chromatographic separation was 205 206 performed by a DB-624 column (Agilent). For each sample, two 10 mL water samples were 207 taken and stored at 4°C. Prior to the analysis, 90 µL of the 0.5 ppm 1,2,4-trimethylbenzene internal standard were added to each 6.9 g sample. Gas samples were collected using a needle 208 209 sampling system controlled by a manometer (1bar/vial). Quantification was performed with 210 methane as reference gas containing 10 mol-ppm of benzene and toluene.

211 2.5- *Treatment of the solid phase*.

Before the experiment, the reservoir rock was rinsed with isopropanol followed by water in order to eliminate potential hydrocarbons and salts such as calcium chloride used during the drilling operation. After rinsing, the solid sample was dried and sterilized overnight in the oven at 90°C. A sample of 450 mL of the solid phase was placed in the basket inside the highpressure reactor. Capillaries used for X-ray tomography were also filled with the solid phase and placed in the basket.

Once the reactor closed, a volume of 60 mL of demineralized water was injected and a flush of N_2 gas was applied. The autoclave was heated to 100 °C during 24h, water vapor was evacuated through the tubes and valves for sterilization. Afterwards the reactor was cooled down, heated to the desired temperature of 36 °C and the water phase (with autochthonous microorganisms) was injected.

After 180 days of experiment at 36 °C and 60 bar, the remaining liquid phase was removed but maintained in sterile and anoxic conditions and was kept at 4°C awaiting further analysis.

Afterwards, the solid basket was quickly and carefully removed from the reactor and stored in a sealed anaerobic jar with anaerobiosis generator and indicator pockets (Dutscher Ref 0260001) before being transported to an anaerobic glove box.

228 2.6- *Thermodynamic modelling*.

The thermodynamic equilibrium at different times of the experiment was calculated using 229 PHREEQC thermodynamic software (Parkhurst and Appelo, 1999). This software allows a 230 231 comprehensive introduction of dissolved species, solid and gaseous phases and the corresponding chemical reactions. The thermodynamic data were obtained from 'Phreeqc' 232 database without any optimization or modification. Phreegc.dat uses the extended Debye-233 Hückel law for activity coefficients and the Redlich-type equation of state for the gas phase. 234 Steady state was assumed for all calculations. All the gas species were considered, but redox 235 reactions of methane and nitrogen were neglected. Water evaporation was considered. Sulfate 236 reduction reaction was not considered in the anoxic part of the experiment. Phreeqc was 237 238 mainly used to calculate gas solubilities in absence of reactions, thus interactions with the 239 solid phase were not considered. The calculations took into account the different water injections occurring during the experiment and the quantity of gas sampled during the 240 chromatography analysis. 241

242 2.7- Cell culture media, strains isolation and brief characterization.

Different synthetic culture media were prepared for the assessment of the microorganisms developed in the reactor just before O_2 addition to the system. In order to prepare these media, a mineral base mimicking the formation water composition was prepared (per liter) by adding 0.0002g NH₄Cl; 0.01g KCl; 0.13g CaCl₂·2H₂O, 0.023g MgCl₂·6H₂O and 0.015g K₂HPO₄, 10 ml of trace elements solution (SL12; (Eichler and Pfennig, 1986)) and 0.5 mL of selenite tungstate solution (Widdel and Bak, 1992). One gram of cysteine as reducing agent and 1 mL

resazurine (0.1% solution, w/v) as redox indicator were supplemented. Depending on the 249 250 targeted metabolic groups, other compounds were also added. For fermentative bacteria, 10g yeast extract; 2.5g glucose; 15g casein peptone were added whereas for sulfate-reducers, 251 0.82g sodium acetate; 0.5g yeast extract; 0.68g sodium formate; 1.1g sodium pyruvate; 2.84g 252 sodium sulfate were added. Media were autoclaved at 110°C for 30 minutes and then cooled 253 254 down at room temperature under a stream of N_2 . At room temperature, 0.2g of NaHCO₃ 255 prepared in anoxic conditions and autoclaved separately from the medium was added to each 256 culture medium. Five milliliters of filtered anoxic V10 vitamins solution (Pfennig et al., 1981) 257 were added. Media were adjusted to a pH of 6.7 and distributed anoxically in Hungate tubes. 258 Enrichment cultures were incubated at 36 °C. The growth was followed by optical density at 600 nm (Spectrophotometer, Spectronic Instruments, Spectronic 401). 259

Several strains were isolated from the last tube showing growth during the Most Probable 260 Number technique. The 15A and Recti_s6 strains were thus isolated from the medium 261 targeting the sulfate-reducers, whereas the 2A strain was isolated from the medium dedicated 262 263 to the fermentative metabolic group. Isolation of each strain was carried out by the method of dilution series in deep agar (Pfennig and Trüper, 1992). Rifampicin was added at a 264 concentration of 10 µg/mL during the Recti_s6 isolation (Koelschbach et al., 2017). The 265 266 purity of the isolated strains was checked under a phase contrast microscope (BX60; Olympus) as well as by sequencing their 16S rDNA genes. For the fermentative strains (2A & 267 268 Recti_S6), searching for the anaerobic fermentation products was conducted using 10 mM of glucose and a N₂ gas phase (1 bar). The gas phase was followed by GC-µTCD (Micro GC 269 Fusion; Chemlys; France). The fermentation products were analyzed using ion 270 271 chromatography (ThermoFisher). This process was conducted at the end of the exponential growth curve, by taking 0.5 mL of culture and filtering it at 0.22 µm (IC Millex®, Millipore) 272 273 in a sterile degassed Hungate tube. Sterile physiological water was added afterwards in order

to reach a final volume of 2 mL. For the sulfate-reducing strain (15A), the experiments were 274 275 conducted in 120 mL penicillin flasks. The strain was cultivated with H₂/CO₂ (80/20, 1 bar) as energy and carbon sources respectively, in addition to 20 mM of sulfate as terminal electron 276 acceptor. The final products obtained from the sulfate consumption by this strain were 277 monitored in these conditions, then with 10 mM of formate as carbon and energy sources and 278 20 mM of sulfate under N_2 atmosphere. The changes in the gas and liquid phases were 279 280 investigated by gas micro-chromatography and ion chromatography, respectively. The microbial growth was followed by optical density at 600 nm as described above. The isolated 281 strains were tested for their oxygen tolerance. Different volumes of air, filtered at 0.1 µm, 282 283 were added to several Bellco tubes. A tube with a nitrogen gas phase and another one containing only air served as controls. The gas phase was analyzed at the beginning and at the 284 end of the experiment by a microthermal conductivity detector (Gc-µTCD; MicroGC Fusion; 285 286 Chemlys; France). The microbial growth was followed by optical density at 600 nm.

287 2.8- Nucleic acids extraction and RNA retro-transcription.

288 At different times of the experiment, samples from the liquid phase were collected in order to co-extract the nucleic acids. The samples were directly filtered under vacuum using a 47 mm 289 PES membrane filters of 0.1 µm porosity (by Sartorius Stedim). Afterwards, the collected 290 291 filters were stored at -80 °C in order to protect the RNA. Filters were crushed in liquid nitrogen and the nucleic acids were extracted using the Fast RNA Prosoil Direct kit (MP 292 293 BIO). DNA was separated from RNA with All Prep RNA/DNA (Qiagen) kit. From the isolated strains, 2 mL of each culture were pelleted by centrifugation at 10.000 g for 5 294 minutes. The DNA was then extracted using DNeasy Power Soil (Qiagen) kit. After 295 extraction, DNA and RNA were quantified using the Quant-itTM dsDNA HS (Invitrogen) and 296 the Quant-itTM RiboGreen (Invitrogen) kits, respectively. The measurements were done using 297 a BioTEK SYNERGY HTX microplate reader. The reverse transcription of the RNA was 298

299 performed *via* retro transcriptase M-MLV (InvitrogenTM) according to the supplier's
300 recommendations.

301 2.9- Polymerase chain reaction and sequencing.

The V3-V4 regions of the 16S rRNA gene of the genomic DNA and cDNA were amplified by 302 nested PCR as previously described (Yu et al., 2015). The search for archaeas was carried out 303 by targeting successively the 21F-958R and 349F-806R regions of the Archean 16S rRNA via 304 305 a nested PCR (Takai and Horikoshi, 2000). The Taq PCR Core Kit (Roche) and the 2720 Thermal Cycler (Applied Biosystems) were used. The bacterial PCR primer pairs (344F-306 307 801R) that were used contained CTTTCCCTACACGACGCTCTTCCGATCT (forward) and GGAGTTCAGACGTGTGCTCTTCCGATC (reverse) adapters. Afterwards, high-throughput 308 sequencing conducted by the GenoToul genomics platform (Toulouse, France), using 309 Illumina's MiSeq 2x250bp technology in accordance with the manufacturer's instructions was 310 undergone. Sequencing data were processed via FROGS analysis pipeline developed by the 311 312 GenoToul genomics plateform in the Galaxy interface (Escudié et al., 2018).

Using the DNA of the isolated strains, the 16S rRNA genes amplifications were conducted
with the primers set 8F-1489R (Lane, 1991; Weisburg et al., 1991). The amplicons were
sequenced by Qiagen Genomic Services (Sanger sequencing; Hilden, Germany).

Raw sequences were archived in public NCBI database (SAMN19579666-SAMN19579701).

The Genbank/ EMBL/ DDBJ accession numbers of the 16S rRNA gene sequences of isolated

318 strains were: 15A (MZ317661), Recti_s6 (MZ317662) and 2A (MZ317660).

319 2.10- Quantitative PCR.

Quantitative PCR (qPCR; Biorad CFX Connect) *via* Takyon NO ROX SYBR 2X MasterMix
blue dTTP (Eurogentec) was used to determine the number of 16S rRNA gene copies present
in each sample during the incubation at high pressure (from DNA and cDNA). The primers

used were 338F (Suzuki and Giovannoni, 1996) and 518R (Muyzer et al., 1993). All reactions
had a 20µL final reaction volume in accordance with the supplier's instructions. The reactions
were denatured at 95°C/3min followed by 40 cycles of 95°C/10sec, 55°C/20sec and
72°C/20sec. The number of copies was calculated starting from a standard with serially 10fold diluted pCRTM 2.1-TOPO plasmid (TOPO TA cloning kit, Invitrogen).

328 2.11- X-Ray Tomography.

329 The morphological alterations of the solid phase were followed by X-ray tomography. Capillaries made of borosilicate with an internal diameter of 2 mm and a height of 3 cm were 330 filled with the solid phase (sand) and positioned in the reactor. The bottom of each capillary 331 was blocked with cotton to prevent loss of material. Before the experiment, capillaries were 332 scanned by X-ray tomography for initial reference. After the completion of the experiment, 333 the capillaries were removed and sealed at both ends with waterproof glue and stored at 4 °C 334 in anaerobic glove-box until scanning. A Zeiss Xradia Versa 510 tomograph was used. Based 335 336 on a prior sensitivity study, the selection of the acquisition parameters was centered on 337 optimizing the image contrast, signal-to-noise ratio and acquisition duration. Typically a set of 1601 radiographies of 1024x1024 pixels with a pixel size of 2.5µm was acquired. The X-338 ray generator was operated at 40 kV and the exposure time of a single radiography is 10s. The 339 340 set of recorded radiographs was reconstructed with XRM Reconstructor® (Zeiss, version 14) in order to obtain a stack of cross-sections forming a digital volume of the sample. This 341 342 reconstruction software was based on the standard filtered back projection algorithm. The data were processed, visualized and interpreted using Dragonfly® (ORS, Version 2020.2.0.11731). 343 A filtering step was first applied in order to improve the signal-to-noise ratio of the data. The 344 345 subsequent segmentation step enabled the extraction of the different phases (solid, liquid, biomass) for further quantitative analyses such as assessment of the proportion and location of 346 each phase in the studied volume and comparison of data sets before and after the experiment. 347

348 2.12- X-Ray Diffraction and Scanning Electron Microscopy.

X-Ray diffraction (XRD) was used to study the mineralogy of the reservoir rock. Mineral 349 identification was performed using a Bruker D2 Phaser powder diffractometer with a Cu Ka 350 radiation source. XRD patterns were recorded over 5° to 90° 2 Θ with a 0.02° step and a 0.5 s 351 counting time per step. Samples from the reservoir rock were analyzed before the experiment. 352 353 Bulk rock was analyzed in order to obtain the global mineral phases and a color-based fractionation was performed in order to identify potential minor crystallized phases. After the 354 completion of the reactor experiment, samples from different locations of the basket were 355 collected in the glove box. The collections were performed from three different depths that 356 were 1 cm away from each other. The samples were then dried up under gaseous N₂ in an 357 358 anaerobic glovebox at ambient temperature. All samples were manually ground and sieved into a homogeneous powder at 100 µm, the product was then analyzed by XRD. Identification 359 of crystallized phases was done using the DIFFRAC.EVA software. 360

Scanning Electron Microscopy (SEM) was used to localize mineral phases and to detect 361 potential formation of biofilm. Collected samples from three different depths, that were 1 cm 362 away from each other, were immediately fixed in a mixture of cacodylate 0.1 M and 2.5 % 363 glutathione pH 7.2. They were then dried out in successive solutions containing 30 to 100 % 364 ethanol and finally evaporated using hexamethyldisilazane (HMDS). Afterwards, the samples 365 366 were mounted on stubs and platinum coated. Observations were carried out using a Gemini SEM 300 (Zeiss) combined to Energy Dispersive Analysis (EDX) at 8 kV in secondary 367 electron mode (SE) or backscattering electron mode (BSE). 368

369

370 3. <u>Results</u>

371 *3.1- Monitoring of the physico-chemical evolution during abiotic experiments.*

After approximatively 10 days of mixing the liquid and the solid phases, water composition 372 373 showed an increase of chloride $(3.05 \pm 0.15 \text{ mM})$, sulfate $(0.29 \pm 0.01 \text{ mM})$ and calcium $(2.18 \pm 0.01 \text{ mM})$ \pm 0.22 mM) concentrations in the water, indicating that the mineral matrix released theses 374 ions. Very limited quantity of acetate appeared in the liquid phase (0.005 ± 0.0003 mM). The 375 solubilities at the thermodynamic equilibrium of the system liquid - solid - gas were 376 calculated according to the measured variations in the liquid phase. Under a total pressure of 377 60 bar, a temperature of 36 °C and with a gas composition of 98.3% CH₄, 0.8% CO₂ and 0.9 378 % O₂ at equilibrium, the calculated solubility of CH₄, CO₂ and O₂ using PhreeqC and 379 phreeqc.dat are $5.96 \cdot 10^{-2}$, $8.88 \cdot 10^{-3}$ and $5.25 \cdot 10^{-4}$ mol.Kgw⁻¹, respectively. 380

381 3.2-Monitoring of the physico-chemical evolution in the liquid phase during biotic
382 experiment.

Throughout the experiment, liquid and gas samples were collected in order to perform the 383 different analyses. The solid phase was analyzed before and after the experiment. The liquid 384 385 phase was composed of the formation water with indigenous microorganisms. As expected 386 from a deep aquifer formation water (- 582 m), the latter is highly reduced with an Eh around -200 mV (Table 1). The ionic water content is dominated by bicarbonate and sulfate amongst 387 anions and calcium amongst cations. The total concentration of dissolved elements in the 388 389 water can be estimated by summing bicarbonate, sulfate, chloride, calcium, magnesium, potassium and sodium knowing that 4.79 mM (286,6 mg.L⁻¹) represents only 0.8 % of 390 391 seawater salinity (Banks et al., 2014). Such low salinity strongly suggests that the formation water is of meteoric origin. The concentrations of dissolved Ca⁺ and HCO₃⁻ confirm an 392 interaction between water and carbonate minerals such as lithoclasts observed in the reservoir 393 394 rock. The origin of sulfate can be hardly depicted and could have resulted from the dissolution of trace amounts of gypsum or anhydrite or from oxidation of sulfides in more crucial 395 conditions. Despite its relatively low concentration, the dominance of sulfate as a terminal 396

397 electron acceptor in the formation water suggests that sulfate-reducing bacteria are the398 dominant microorganisms in such ecosystem.

One month before its injection into the reactor, the formation water (downhole sample) 399 400 containing the microorganisms, stored until now at 4°C, was supplemented with wellhead water, sterilized by 0.1µm-filtration and maintained at room temperature. At the beginning of 401 the biotic experiment, sulfate concentration was $9.20 \cdot 10^{-2} \pm 4.60 \cdot 10^{-3}$ mM (*i.e.* $1.28 \cdot 10^{-1} \pm$ 402 $6.40 \cdot 10^{-3}$ mmol; Figure 3a). During the experiment, analytical measurements showed an 403 initial decrease in the amount of sulfate after 10 days from the start, followed by its complete 404 405 consumption on day 32. In order to maintain the dynamic equilibrium, formation water 406 supplemented with sulfate was added. The whole added formation water was filter-sterilized with a 0.1 µm filter to avoid any microbial interference. Three successive water injections 407 were performed on days 48, 104 and 126, respectively. After each injection, an expected 408 increase in sulfate quantity was noted. Sulfate quantity increased by $9.62 \cdot 10^{-2} \pm 4.81 \cdot 10^{-3}$ 409 mmol, $5.10 \cdot 10^{-2} \pm 2.55 \cdot 10^{-3}$ mmol and $1.50 \cdot 10^{-2} \pm 7.50 \cdot 10^{-4}$ mmol on days 50, 105 and 131, 410 respectively. Following these increases and before O₂ injection, sulfate consumption resumed 411 and continued until sulfate's complete depletion from the system. During the first 130 days, 412 $2.61 \cdot 10^{-1} \pm 1.31 \cdot 10^{-2}$ mmol of sulfate were consumed. The third injection was applied to refill 413 414 the sulfate quantity in order to replenish the system stores before O₂ injection.

During this first period of the experiment, sulfate quantity showed a notable progressive reduction. However, after oxygen injection into the system, the previously noted progressive reduction in sulfate quantity was noted to subside and even reverse into a remarkable upsurge. This estimated increase in sulfate quantity was of $4.30 \cdot 10^{-2} \pm 2.15 \cdot 10^{-3}$ mmol. Using ionic chromatography coupled with a variable wavelength detector, no sulfide ions were detected in the liquid phase. Similarly no hydrogen sulfide was detected in the gas phase by gas chromatography (1 ppm as minimum detection limit). Taking into account the amount of sulfate depleted from the system, sulfate-reducers were expected to produce $2.61 \cdot 10^{-1} \pm 1.31 \cdot 10^{-2}$ mmol of sulfide, bisulfide or hydrogen sulfide at most. Even if the sulfur was supposedly only in the form of hydrogen sulfide, it constituted less than 0.006 % of the total gas quantity. The value is too low to be detected by gas chromatography.

At the beginning of the experiment and simultaneously with sulfate consumption, an increase in acetate quantity was noted and is shown in Figure 3b. Acetate was produced in the reactor during the first 45 days despite not being detected initially in the formation water. Its quantity increased from zero to $1.23 \pm 6 \cdot 10^{-2}$ mmol at day 13, then attained $2.09 \pm 1 \cdot 10^{-1}$ mmol at day 50 with an additional increase until reaching a peak of $2.34 \pm 1 \cdot 10^{-1}$ mmol. This quantity remained stable until the end of the incubation.

Similarly to *in situ* conditions, mono-aromatic hydrocarbons including toluene and benzene 432 433 were supplemented in the gas phase. Monitoring of these compounds was carried out in the liquid phase. At day 8, the molar quantity of toluene was around $2.55 \cdot 10^{-3} \pm 1 \cdot 10^{-4}$ mmol 434 (Figure3c), which decreased after 35 days before increasing again after toluene injection. 435 After this injection, $2.08 \cdot 10^{-3} \pm 1 \cdot 10^{-4}$ mmol were present in the system at day 49. Afterwards, 436 the toluene gradually decreased to $4.12 \cdot 10^{-4} \pm 6 \cdot 10^{-5}$ mmol at the end of the experiment (day 437 179). During 180 days of incubation including, 130 days in anaerobiosis and 50 days in 438 439 aerobiosis, no benzene degradation was observed in these conditions.

440

441 *3.3- Monitoring of the physico-chemical evolution in the gas phase during biotic experiment.*

442 Methane is the most abundant component in the gas phase. At the beginning of the 443 experiment, 60.4 ± 0.6 bar of the gas mixture CH₄ + 1% CO₂ was added to the system at a 444 temperature of 36 °C. Considering the compressibility factor of 0.9 which is calculated by the 445 PhreeqC model based on Peng-Robinson's equation, $4.52 \pm 2 \cdot 10^{-1}$ mol of CH₄ were injected.

At day 4, $5.99 \cdot 10^{-2} \pm 2x10^{-3}$ mol of CH₄ were lost from the gas phase (Figure 3d). Gas 446 447 quantities were calculated based on the total pressure measured by the barometer and the gas phase composition analyzed by the gas chromatography. The calculation performed via 448 Phreeqc indicated a solubility of 6.25 · 10⁻² mol.Kgw⁻¹ in this experiment's conditions. Thus, 449 taking into account the uncertainty, the CH₄ quantity lost from the gas phase corresponded 450 entirely to its soluble equivalent in the liquid phase. In order to readjust the total pressure after 451 stabilizing the piston position, a second and last injection of the gas mixture was performed 452 on day 6. This injection induced a $2.16 \cdot 10^{-1} \pm 1 \cdot 10^{-2}$ mol increase in methane quantity. After 453 each water injection, a certain number of methane moles transfer to the liquid phase in order 454 to preserve the gas-unsaturated liquid equilibrium. The quantities transferred to the formation 455 water were $6.27 \cdot 10^{-2} \pm 3 \cdot 10^{-3}$ mol, $7.77 \cdot 10^{-2} \pm 3 \cdot 10^{-3}$ mol and $7.13 \cdot 10^{-3} \pm 3 \cdot 10^{-4}$ mol after the 456 first, second and third water injections, respectively. Afterwards, no variations in methane 457 458 quantity were observed until the end of the experiment. The curve in the Figure 3d represents the variations in methane quantity throughout the experiment based on the thermodynamic 459 460 model. This model fits well the experimental data obtained during the study.

Carbon dioxide was present in an initial proportion of 1 % of the gas mixture, corresponding 461 to an initial quantity of $4.56 \cdot 10^{-2} \pm 2 \cdot 10^{-3}$ mol in the gas phase. Figure 3e showed the 462 variations of CO₂ quantity during the experiment. Similarly, to methane, an initial decrease in 463 CO₂ quantity was observed due to the gas-liquid equilibrium. Following this phase, the 464 CH₄/CO₂ injection applied to readjust the pressure generated an increase in CO₂ quantity of 465 $2.34 \cdot 10^{-4} \pm 1 \cdot 10^{-5}$ mol. Throughout the experiment, a decrease in CO₂ quantity was observed. 466 The last measurement that was done on day 178 showed $1.56 \cdot 10^{-2} \pm 7 \cdot 10^{-4}$ mol of CO₂ in the 467 gas phase. The curve representing the thermodynamic variations of CO₂ in the gas phase was 468 presented in the same figure. The model showed that the thermodynamic equilibrium and 469 470 gases solubility were reached before the reinjection of CH₄/CO₂. The reinjection did not have 471 any significant impact on the thermodynamic equilibrium in view of the minimal quantity that 472 was added. The calculated pH was estimated at 5.1. However, the continuous decrease in the 473 carbon dioxide quantity as well as the remarkable difference between the model and the 474 experimental measurements cannot solely be explained by thermodynamic calculations and 475 suggested the presence of an autotrophic metabolism and the incorporation of the carbon 476 inside the biomass.

On day 130 of incubation, O₂ was injected in the reactor. At day 131, $4.16 \cdot 10^{-2} \pm 2 \cdot 10^{-3}$ mol 477 were quantified, constituting 1% of the gas phase (Figure 3e). The quantity of oxygen is 478 calculated based on the total pressure of the reactor with a compressibility factor of 0.9 and 479 480 the molar fraction of O_2 analyzed by the gas chromatography. At the end of the experiment (day 178), $3.57 \cdot 10^{-2} \pm 1 \cdot 10^{-3}$ mol of O₂ remained in the gas mixture. Noting the increase in 481 sulfate in the liquid phase after O_2 injection, 8.60 x10⁻² mmol of O_2 had potentially oxidized 482 sulfide into sulfate. Calculated variations of O₂, taking into account the number of moles lost 483 due to analysis, are represented in the curve. Note that the difference between the 484 485 experimental and calculated values could be due to potential mineral oxidation that occurred in the reactor such as sulfur oxidation. 486

As previously explained, the experiment aims at studying the effects of the 487 488 CH₄ -1% CO₂ -1% O₂ mixture in a deep storage aquifer. Hydrogen was not present in the injected gas mixture. However, limited quantities of H₂ near detection threshold were detected 489 between day 4 and day 45 of the experiment as shown in Figure 3f. Before its depletion, the 490 maximum measured quantity at day 31 was estimated at $2.51 \cdot 10^{-4} \pm 1 \cdot 10^{-5}$ mol, which 491 represented 0.005% of the gas phase. In fact, this amount is too low to be considered at that 492 time of the experiment. However, it is worth mentioning that the peak of H₂ was observed at a 493 period when sulfate was undetected in the system. Surprisingly, H₂ accumulation clearly 494 occurred after O₂ injection into the system. The amount of H₂ over time was shown through a 495

496 positive slope that started one day after oxygen injection and continued until the end of the 497 experiment. The last measurement of H₂ in the gas phase was $1.99 \cdot 10^{-2} \pm 9 \cdot 10^{-4}$ mol, which 498 represented 0.5% of the gas phase. The experimental H₂ values indicated continuous 499 production of this gas. Considering a solubility of $2 \cdot 10^{-4}$ mol.Kgw⁻¹ (*via* PhreeqC), $1.75 \cdot 10^{-4}$ 500 mol of H₂ were present in the liquid phase at the end of the experiment.

501 *3.4- Evolution of the bacterial community during the experiment.*

502 Besides the importance of identifying O₂ effect on microbial life, analyzing the qualitative 503 and quantitative variations in microbial diversity throughout the experiment can help understanding the changes in the liquid and the gas phases. The presence of archaea could not 504 505 be demonstrated by PCR and nested-PCR approaches targeting this taxonomic group. The bacterial diversity results based on the sequencing of bacterial 16S rRNA genes were 506 507 combined with the results of quantification of these same genes obtained by quantitative PCR in each sample (Figure 4). At the beginning of the experiment, the microbial community of 508 $2.1 \cdot 10^5 \pm 1.3 \cdot 10^5$ cells/mL stored since five months at ambient pressure at 4 °C was 509 510 dominated by Geobacteraceae > Desulfovibrionaceae > Rhodocyclaceae > Spirochaetaceae. 511 The proportions of each of the aforementioned strain species were noted to vary during the first 43 days of incubation. On one hand a reduction in Geobacteraceae and 512 513 Desulfovibrionaceae was observed, in contrast to a growth in Spirochaetaceae and Peptococcaceae. This new equilibrium in the community was relatively stable until 514 515 incubation day 105 when the Peptococcaceae concentration was noted to be increasing. Peptococcaceae growth coincided with the reduction in toluene in the liquid phase (Figure 516 517 3c). The oxygen injection had a significant impact on the microbial community presented by a substantial reduction in the bacterial cell concentration from $1.1 \cdot 10^5 \pm 2.0 \cdot 10^4$ cells/mL (t₁₃₀) 518 to $3.7 \cdot 10^3 \pm 1.2 \cdot 10^3$ cell/mL (t₁₃₃) reaching an endpoint of $4.1 \cdot 10^2 \pm 8.4 \cdot 10^1$ cell/mL (t₁₇₉). 519 Desulfovibrionaceae and Sphingomonodaceae retained a detectable presence despite having 520

521 higher concentrations prior to oxygen injection. Members of the *Sphingomonodaceae* may 522 have aerobic metabolism in addition to having fermentation capacities allowing them to 523 produce H_2 by degrading organic molecules.

524 Via the 16S rRNA gene, the bacterial taxonomic diversity of active cells was monitored and the results are presented in the Figure 5 and Table 2. The bacterial strains' activity seems to 525 be restricted and different from that seen at atmospheric pressure (intensity of activity at day 0 526 527 compared to the activity during the incubation). Thus, the Geobacteraceae (blue), 528 Rhizobiaceae (brown), Desulfovibrionaceae (black yellow, below) and Spirochaetaceae (dark blue) were the most active in the original community obtained from the sampler that evolved 529 530 for several months at atmospheric pressure in anaerobiosis. Just before O_2 injection (day 130), activities of Desulfovibrionaceae (yellow), *Peptococcaceae* 531 strong (green) and Thermodesulfovibrionaceae (blue) were detected. On day 179 at the end of the incubation and 532 even though their microbial activity were quite low, Spirochaetaceae 533 and Desulfovibrionaceae strains could still be detected. Hence, the sulfate-reducing metabolic 534 535 group had suffered but has not completely disappeared.

In addition, the liquid phase was sampled prior to O₂ injection in order to target and cultivate 536 sulfate-reducers and fermenters in special media, as those microorganisms are considered key 537 538 players in deep aquifers. From the sulfate-reducers' medium, the strain 15A, a representative of the Desulfovibrio vulgaris strains, was isolated and identified (99.72% identity with 16S 539 rDNA sequence of D. vulgaris strain "Miyazaki F"). The strain 15A had a strictly anaerobic 540 541 growth. It was capable of growth with CO₂ and H₂ as sole carbon and energy sources. Regarding its heterotrophic growth, formate oxidization was tested and turned to be 542 543 incomplete with noted production of acetate. As expected, the strain 15A did not produce acetate when grown autotrophically. Two fermenters were isolated, strain 2A and strain 544 Recti_s6. Strain 2A is phylogenetically close to Pleomorphomonas koreensis (99.70% of 545

identity; 16S rDNA). Grown on glucose-based medium, strain 2A produces formate, acetate, 546 547 CO₂ and H₂ even in oxic conditions such as the air in the head space. Strain Recti s6 is related to Rectinema cohabitans strain HM (99.83% of identity) which a spirochaete isolated 548 from an anaerobic naphthalene-degrading enrichment culture (Koelschbach et al., 2017). 549 Finally, representatives of the *Bacillus* genus were also detected through the identification of 550 their characteristic structures and spore forms (data not shown). These bacteria are strictly 551 552 aerobic and are able to maintain under anoxic conditions through their ability to sporulate. Spores protect these bacterial until more favorable conditions for germination occur. 553

554 *3.5-* Evolution of the solid phase during the experiment.

XRD was performed on the reservoir rock that is sampled from the deep aquifer and 555 introduced in the reactor and on the solid sample after the experiment. At the end of the 556 experiment, X-ray diffraction performed on the collected rock did not demonstrate any 557 significant changes in the solid crystallized phases before and after the experiment (Figure 6). 558 559 Mainly composed of quartz with minor calcite components, the solid sample did not vary 560 significantly after the experiment. However parallel measurements were performed on a reservoir rock sample that was immerged in the formation water containing the microbial life 561 with the substitution of CH₄-CO₂-O₂ for N₂ gas. These measurements indicated the formation 562 563 of iron sulfide (Figure 6). Thus the possibility of iron sulfides production during the experiment with its subsequent oxidation after O₂ injection cannot be excluded. SEM-EDX 564 565 showed high amount of clay minerals in between quartz grains, and Fe and S colocalization 566 suggested the presence of iron sulfides in agreement with XRD (Figure 7). No variations was observed by depth in the basket and no biofilm could be detected in these conditions. 567

The average porosity obtained by means of X-ray tomography from four different capillaries did not show any significant variation with an initial porosity of around 43.5 % and a final one of around 44.5% with an estimated uncertainty of ± 2 %. Thus, no recrystallization or

potential biofilm could be demonstrated in agreement with mineralogical observations. Note
that black particles were observed in the solid phase at the end of the experiment. Similar
black particles produced during a bio-geochemical interactions study were identified as pyrite.

574

575 4. Discussion

576 The role and contribution of green and renewable energy in the energy mix is more than ever 577 required in order to counteract and reduce the notorious effects of greenhouse gas emissions. As part of energy transition, natural gas will become greener through the production and 578 579 storage of biomethane. However, despite that biomethane is an attractive source of renewable energy, the effects of its storage particularly in massive amounts remains unknown. Thus, 580 identification and comprehension of associated phenomena to biomethane storage is 581 fundamental for the basis of any territory's energy strategy. In order to secure energy 582 supplies, to face consumption peaks and variations between summer and winter, a significant 583 584 part of natural gas can be stored in geological structures such as salt cavities or deep aquifers. That said, biomethane could replace natural gas on the long term and can be stored within 585 these geological structures. Unlike natural gas, O₂ is added to biomethane during sulfide 586 reoxidation process before injection into the gas network. To our knowledge, this is the first 587 study to evaluate the potential impact of O₂ in anoxic ecosystems such as deep aquifers in this 588 setting. The scenario tested in this study is the most unfavorable case of biomethane storage 589 consisting of an O₂ supply of 1% in the context of an Eocene aquifer which is mainly 590 composed of sands, sandstones and carbonate deposits (Brenot et al., 2015). Based on the X-591 ray diffraction results, calcite, iron sulfides (pyrite and marcasite), oxi-hydroxides (maghemite 592 593 and hematite) and clay minerals were detected in traces. The monitoring well Ab L 1 sampled for this study is the closest to the natural gas bubble (Ranchou-Peyruse et al., 2019). 594 At an approximate depth of 600 m, interactions between the formation water and the injected 595

gas are at their peak. Under a high pressure of 60 bar, oxygen solubility is more important 596 597 than its solubility at the surface even at a 1% proportion. The Ab L 1 formation water has a very low salinity of 4.79 mM (286.56 mg.L⁻¹) which is equivalent to only 0.8 % of marine 598 599 salinity (Table 1). However the abiotic experiment showed salt release from the solid into the liquid phase despite rinsing of the rock during the preparation. Calcium and chloride were the 600 601 main ions detected. Calcium chloride is a brine known to be used during drilling operations. 602 Nevertheless, stoichiometry suggests another origin for calcium which could be released from 603 the calcite detected in the reservoir rock. Even in small quantities of few mmoles, changes in the liquid's salinity have the potential to alter the thermodynamic equilibrium. 604

605 *4.1- Sampling and pre-incubation.*

606 The formation water containing indigenous microorganisms was sampled 5 months before the start of the high-pressure biotic experiment. Water was collected using a downhole sampler 607 ensuring the purity of the microbiological sample and maintaining it under high-pressure until 608 609 reaching the laboratory. Under controlled conditions, the pressure was gradually lowered in 610 order to avoid damaging or destroying the microorganisms through rapid depressurization. The formation water was then stored at ambient temperature and under anoxic conditions with 611 N₂ as gas phase. One month before the experiment, the formation water with the 612 613 microorganisms was supplemented with sterile water from the same study site in order to 614 reactivate the microorganisms. During this pre-incubation period, the microorganisms already 615 started to consume different elements present in the water particularly sulfate. Once injected 616 in the reactor, liquid phase was supplemented with sulfate by the release of these ions from the rock. Sulfate ions present in the solid phase may be due to pyrite oxidation or to drilling 617 618 brines. Incubation at atmospheric pressure seems to have particularly favored the Geobacteraceae family, which proportion drastically decreased in the pressurized reactor. 619 Similar changes affected the Desulfovibrionaceae family which are sulfate-reducing 620

621 proteobacteria previously described in this site (Ranchou-Peyruse et al., 2019). These findings 622 encourage undergoing further studies assessing the behavior of microbial communities in 623 deep environments under high pressures, even in continental undergrounds where pressures 624 are known to be lower than deep marine ecosystems.

625 *4.2- Physicochemical and bacterial evolution before oxygen injection.*

The objective of the custom-designed reactor in this study is to simulate as precisely as 626 627 possible a gas storage aquifer. Anoxic conditions were applied from the beginning of the experiment with a gas phase composed of CH₄ with 1% of CO₂ and small quantities of 628 benzene and toluene. The entire rock was initially immersed in the formation water during the 629 first 5 days in order to colonize it with microorganisms. Afterwards the water level was 630 lowered allowing the upper part of the rock to bath in the gas phase whereas its lower part 631 632 remained in the liquid phase with the microorganisms. Microbial communities were able to grow under a pressure of 60 bar and a temperature of 36 °C, similarly to the aquifer's in situ 633 634 conditions. The thermodynamic modeling studies helped to identify the equilibrium status of 635 our system. The parameters such as the gas quantity and volume, the aqueous phase composition and quantity and the temperature of the system were fixed. PhreeqC model 636 calculated the solubility of each gas in the aqueous phase under the system's conditions, and 637 638 the quantity of each gas in the gas phase at equilibrium. At day 5 and before the reinjection of CH₄+CO₂, the quantities of CH₄ (4.46 $\pm \cdot 2.2 \cdot 10^{-1}$ mol) and CO₂ (2.97 $\cdot 10^{-2} \pm 1.5 \cdot 10^{-3}$ mol) in 639 the gas phase were equal to those calculated by PhreeqC (4.41 and $3.37 \cdot 10^{-2}$ mol), 640 respectively. These findings confirm that the thermodynamic equilibrium was attained on day 641 5. The re-injection of the gas mixture on day 6 inducing a small increase in the quantities of 642 643 these gases (Figures 2d and 2e) did not cause a significant difference in the thermodynamic equilibrium. On day 20, the reduction in CO_2 quantity from the gas phase was greater than the 644 value calculated at the thermodynamic equilibrium suggesting an additional consumption 645

which can only be explained by integration of CO_2 in the bacterial biomass. The detection of 646 647 H₂ followed by its consumption supports this hypothesis. Carbon dioxide and hydrogen are consumed by autotrophic microorganisms that use H₂ as a source of energy and electrons. The 648 absence of archaea in the microbial community confirms the lack of CO₂/H₂ consumption via 649 methanogenesis in this study. Since the beginning of the experiment, sulfate reduction was 650 651 noted in the liquid phase in the absence of O₂. That observation necessitated three sulfate-652 supplemented formation water injections in order to maintain the incubation. The bacterial community analysis confirmed the presence of the Desulfovibrionaceae family (Figure 4). 653 Moreover, the 16S rRNA transcripts studies confirmed their activity in the original 654 655 community (Figure 5). Previous studies on the Ab_L_1 site recognized the activity of the sulfate-reducing bacteria, particularly the delta-proteobacteria which include the 656 Desulfovibrionaceae family (Ranchou-Peyruse et al., 2019). We conclude that the anaerobic 657 658 activity of sulfate-reducing bacteria was responsible for the depletion of sulfate from the system. The Desulfovibrio vulgaris 15A strain isolated in this study showed an ability to 659 produce acetate under heterotrophic conditions. Acetate level was noted to increase during the 660 first month of incubation and remained stable until the end of the experiment suggesting a 661 simultaneous absence of acetoclastic bacteria (Figure 3b). Moreover, acetate was not among 662 663 the ions released from the solid phase during the abiotic experiments. Acetate could have also been produced by fermentative bacteria present in the community such as isolated 664 Pleomophomonas koreensis 2A and Rectinema cohabitans Recti_s6 strains. Under 665 fermentative conditions, these two strains have been shown to produce formate, acetate, CO2 666 and H₂. Therefore we conclude that in the first days of incubation, there was a significant 667 activity of fermenters particularly Spirochaetaceae and heterotrophic sulfate-reducers. The 668 latter reduce sulfate from organic molecules present in the water, mineral matrix but also in 669 the necromass. It is also evident that the H₂ produced during fermentation was directly 670

consumed by the sulfate-reducers including the Desulfovibrio vulgaris 15A strain, as sulfate 671 672 depletion coincided with H₂ detection. The hydrogenotrophic ability of certain strains of the Desulfovibrio genus has been previously reported (Bagnoud et al., 2016; Detmers et al., 673 2004). Conversely, sulfate supplementation resulted in H₂ disappearance. Similarly, the 674 reduction in CO₂ (Figure 3e) could be partially explained by this autotrophic growth that uses 675 676 this carbon source for its biomass and metabolism. However the sulfide compounds produced through the sulfate-reduction metabolism were not detected in the liquid nor the gas phases. 677 The black particles observed in the solid basket and the reactor at the end of the experiment, 678 identified as pyrite by XRD, suggest that iron sulfide was formed by the interaction of ferrous 679 680 irons present in the reservoir rock and the formation water with the produced sulfide compounds (Berlendis et al., 2010; Caldwell et al., 2008; Fry et al., 1997; Park et al., 2006; 681 Zettlitzer et al., 2010). 682

683 *4.3- Monoaromatic hydrocarbons biodegradation.*

684 Several studies have demonstrated the potential for biodegradation of monoaromatic 685 hydrocarbons by autochthonous microbial communities in deep aquifers used for natural gas storage (Berlendis et al., 2010; Aüllo et al., 2016; Ranchou-Peyruse et al., 2017). In this study, 686 the gas phase was supplemented with benzene and toluene at concentrations similar to those 687 688 typically found in natural gas (7.95 ppm of benzene and 3.57 ppm of toluene). A reduction in toluene started around day 100 until the injection of O₂. No toluene reduction was noted in 689 690 oxic conditions. A growth of the Peptococcaceae strains (Figures 4 and 5) can be noted in the same period. This bacterial family includes gram-positive sulfate-reducing and fermentative 691 bacteria that have previously been associated with biodegradation of monoaromatic 692 693 hydrocarbons in similar environments (Abu Laban et al., 2010; Aüllo et al., 2016; Kleinsteuber et al., 2008; Liu et al., 2004; Morasch et al., 2004; Ranchou-Peyruse et al., 2017; 694 Robertson et al., 2000; Taubert et al., 2012). This hypothesis is consistent with the current 695

level of understanding of the behavior of microbial communities at the gas bubble / formation 696 697 water interface. In close proximity to the gas bubble, exogenous molecules specifically carbon molecules transferred from the gas to the formation water become more accessible to the 698 microorganisms. This zone that is less oligotrophic than the rest of the aquifer appears to be 699 more favorable for the development of microorganisms such as Desulfovibrio. The 700 Desulfovibrio genus is known to be more competitive and faster growing than 701 702 Desulfotomaculum (Peptococcaceae) when present in environments with abundant organic 703 molecules (Aüllo et al., 2013). After O₂ injection, a notable decline in sulfate reduction, microbial activity and toluene degradation were noted. Terminating the natural bio-704 attenuation of aromatic hydrocarbons could be a side effect of massive O2 influx into UGS 705 and should not be neglected in the management of these geological storages. 706

707 *4.4- Evolution of the four-phase system after oxygen injection.*

Once the O₂ was injected in the gas phase and based on abiotic physicochemical reactions, the 708 simulation model predict that the value at saturation $(4.42 \cdot 10^{-4} \text{ mol.Kgw}^{-1})$ in the liquid phase 709 was experimentally reached 5 days after the O2 injection (i.e., t136). However, the 710 experimental data showed a constant decrease in the gas phase of O2 until the end of the 711 experiment. This reduction is linked to: i) the solubilization of O_2 in the liquid phase (0.67) 712 713 mmol estimated from the model); ii) the reoxidation of sulfide into sulfate (8.60 $\times 10^{-2}$ mmol), the sulfide potentially trapped in the form of iron sulfides in the rock matrix (e.g. pyrite). 714 715 Many studies showed the immediate underground oxidation of pyrite after O₂ injection in the storage systems (Guo et al., 2021; Kushnir et al., 2012; Lu et al., 2016; Vu et al., 2018) 716 causing an increase in sulfate quantities (Yuan et al., 2009); iii) and a possible reoxidation of 717 718 other reduced elements, such as ferrous ions. The decrease in O₂ quantity was more significant and may be due to different gas-solid interactions not identified during our study. 719 The solubility of O_2 simulated by the model can be estimated to be around $4.42 \cdot 10^{-4}$ 720

mol.Kgw⁻¹ in the liquid phase during the last 50 days of the experiment. If a comparable water 721 is placed under different environment and conditions such as air and 36 °C temperature, O₂ 722 saturation would be of around 2.1.10⁻⁴ mol.Kgw⁻¹ at equilibrium (Benson and Krause, 1980; 723 Millero et al., 2002) and an aerobic microbial ecosystem would develop (Stiff et al., 1992). 724 The hyperoxic water in our reactor was clearly toxic to the strictly anaerobic microorganisms, 725 726 as evidenced by the drastic decrease in bacterial taxonomic diversity and its activity from the 727 day following O₂ injection (Figures 4 and 5) even if the existence of some anoxic microniches could be supposed. It is probable that part of O₂ could have reacted with the organic 728 matter resulting from necromass. 729

730 Some microorganisms survived these new environmental conditions and retained a metabolic activity as demonstrated by a surprising constant production of H_2 following O_2 injection that 731 persists till the end of the experiment. Analyses performed on fermentative strains 2A and 732 Recti_s6 demonstrated their capability to produce H₂ in oxic conditions. With the reduction in 733 734 the majority of strict anaerobic bacteria and the decline in hydrogenotrophic sulfate-reducers 735 such as *Desulfovibrio vulgaris* 15A, sulfate and H₂ were no longer consumed, explaining its 736 accumulation in the gas phase. Despite that some strains of the *Desulfovibrionaceae* family were able to survive O₂, their activity is significantly slow and even negligible. A study 737 738 conducted by Yuan et al. showed the occurrence of sulfate-reduction in anaerobic environment after undergoing phases of oxygen stress. Sulfate-reducing bacteria can have a 739 740 relative temporary resistance to oxic environments (Yuan et al., 2009) which explains the detection of their low activity. It is possible that some aerobic microorganisms such as 741 Bacillus strains and/or aero-anaerobic microorganisms such as Sphingomonadaceae may have 742 consumed some of the O₂ as well (Kosako et al., 2000). 743

744 Despite that the solid phase demonstrated no changes in porosity % or mineral composition 745 among others, the pre and post experimental analyses cannot completely rule out the

possibility of potential changes secondary to O₂ injection in the UGS. In fact, the reservoir 746 747 rock sample was in contact with O₂ present in the air before the experiment and was scanned by X-ray tomography and XRD in oxic conditions. During the experiment, interactions 748 749 occurred in anoxic environment and produce minerals such as metal sulfides. Following O2 injection, changes such as mineral oxidation occurred. At this stage, the solid phase was in 750 751 oxic conditions, similarly to its state before the beginning of the experiment. There were no 752 signs suggesting that the metal sulfides precipitates could have significantly reduced the 753 porosity of the rock prior to oxygen injection. However considering these facts, further studies 754 are required in order to assess the state of the solid phase after incubation and before O₂ 755 injection in the system.

During the multidisciplinary experiment, oxygen effects on the different phases of an aquifer 756 during biomethane storage were assessed. Oxygen injection induced mineral oxidation to the 757 reservoir rock (Lu et al., 2016; Vu et al., 2018), especially iron sulfide oxidation causing 758 759 increase in sulfate concentration in the aqueous phase. In fact, iron sulfide oxidation was 760 already observed during CAES (Guo et al., 2021; Jung et al., 2013) and CCS (Jung et al., 761 2013; Shao et al., 2014; Wei et al., 2015) where oxygen was injected as impurity. On a larger 762 scale, reaction of minerals and their dissolution in the aqueous phase can alter the reservoir 763 rock's porosity and therefore affect the storage volume. Another clear result obtained was the deleterious effect of oxygen on the anaerobic microbial communities. Many studies were 764 conducted to confirm the capability of these communities to underground biodegrade 765 766 hydrocarbons, especially the sulfate-reducing bacteria (Aüllo et al., 2013; Ranchou-Peyruse et 767 al., 2019). Even though the microbial concentration drastically decreased after O_2 injection, 768 some of the communities were able to survive. In order to develop, those communities might adapt to the new aquifer conditions. To get their needs of carbon, microorganisms might 769 770 develop capabilities for hydrocarbon biodegradation and therefore, the microbial community would adapt to their new environment. During CCS or biomethane storage in deep aquifers,
oxygen's effects should be considered on many levels depending on the storage conditions
such as the reservoir rock's composition and the microbial communities harbored *in-situ*.

774 5. Conclusion

775 In this experiment, we were able to simulate the aquifer Ab_L_1 situated in South Aquitaine, Southwest France in a reactor. Through a multidisciplinary approach, the impact of 1% of O₂ 776 777 under 60 bar total pressure was evaluated in a UGS aquifer for the first time ever. The relatively simple composition of the rock matrix at the Ab_L_1 site mainly containing sand 778 and to a lesser degree calcite did not seem to react after the addition of O2 at low 779 concentration. Reoxidation of metal sulfide precipitates was expected with subsequent 780 minimal increase in the porosity of the aquifer rock. As of the early 1990s, sulfate was noted 781 782 to decrease in the Ab_L_1 site as a result of the sulfate-reducing activity amplified by the proximity of the gas bubble and by the consumption of organic molecules dissolved in water 783 by sulfate-reducers (1992: ≈18 mg/L SO₄²⁻, 2013: <7 mg/L SO₄²⁻;(Ranchou-Peyruse et al., 784 785 2019)). The reoxidation of metal sulfides would therefore contribute to the restoration of higher sulfate concentrations. Similarly the direct effect on the gases stored in the aquifer 786 (99% CH₄, 1% CO₂) seemed negligible and even absent. Conversely, O₂ had a deleterious 787 788 effect on the strictly anaerobic microbial community interacting with the UGS. This effect manifested as a decline and even a complete arrest of the natural bioattenuation of 789 790 monoaromatic hydrocarbons, a decrease in oxygen amount and finally a surprising H₂ accumulation in the gas phase deriving from fermentative bacteria able to survive the new 791 oxic conditions. 792

The different biological, geological and chemical impacts of biomethane storage containing 1% of O_2 as impurity were evaluated in this study. It will be necessary to carry out new studies with lower oxygen concentrations and for other storage conditions (temperature,

pressure, rocks). Nevertheless, in view of these first results, the elimination of oxygen from 796 797 the gas mixture before its injection must be considered. However oxygen is widely used 798 during different stages of the processing of various other resources such as in CCS or CAES 799 in geological sites. The findings noted in our study shed a new light on O₂ gas effects on the underground environment which is an important step in the green energy era. Such 800 801 observations lay the cornerstone for the future selection of accessible, secure and stable ways 802 of biomethane underground storage which is crucial for building a successful global clean energy strategy. Further studies involving the use of various amounts of oxygen during 803 804 storage of different gases compositions and their subsequent implications are required for a 805 better understanding of the static and dynamic interactions occurring during gas storage 806 practices.

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813 Author Contributions Statement

FC, MRP, PC and ARP co-conceived the study. PH, JM, FC, PC carried out the high pressure
and simulation experiments. MG, MRP and ARP applied the microbiological approaches. PS,
M-PI, IS, PM and GH used the imaging and mineralogical characterization. ML and ILH
analyzed the hydrocarbons. All authors contributed in the interpretation of results and paper
writing.

819 **Declaration of interests**

- 820 DD and PCh are employed by two French companies specialized in geological natural gas
- 821 storage which are Storengy and Teréga respectively.

822 **Conflict of interest**

823 No conflict of interest declared in relation to the work described.

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1042 Tables and Figures

Table 1 - Physico-chemical parameters and compounds of the formation water sampled from
the deep aquifer (Ab_L_1 site), analyzed at atmospheric pressure.

1045 Table 2- Standard deviation related to the 16S rRNA transcripts

Figure 1 - Scheme of the experimental apparatus used during the experiment.

1047 Figure 2- Diagram representing the different steps and analyses undergone though the biotic1048 experiment.

1049 Figure 3- Results of the liquid and gas analyses during the biotic high pressure experiment (60 bar). The vertical line at day 6 represents the CH₄+CO₂ reinjection. The vertical dotted 1050 1051 lines at days 48, 104 and 126 represent sulfate injections. The vertical line at day 132 1052 represents the oxygen injection. a) Sulfate variation in the liquid phase; b) Acetate variation in 1053 the liquid phase; c) Toluene (represented by the blue squares) and benzene (represented by the red triangles) variations in the liquid phase. The vertical line at day 47 represents toluene 1054 1055 injection. d) Methane variation in the gas phase. The red dots and line represent the calculated values; e) Carbon dioxide (represented by the green squares) and oxygen (represented by the 1056 1057 red triangles) variations in the gas phase). The lines represent the calculated values; f) Hydrogen variation in the gas phase. 1058

Figure 4- Evolution of the bacterial community during the incubation at 60 bar, 36° C and with a gas mixture (CH₄, 1%CO₂, 7.95ppm of benzene and 3.57 ppm of toluene) before and after oxygen injection (1%). The taxonomic diversity is monitored by 16S rRNA genes sequencing and the quantification is estimated by qPCR of 16S rDNA gene copies for each ml of sampled water through the experiment. 3a) Monitoring over the 180 days of the experiment. Following the oxygen injection, the quantity of microorganisms has drastically decreased. In order to observe the taxonomic diversity, the boxed part is zoomed in figure 3b. 1066 The variation of the sulfate quantity in the liquid phase is represented by the orange dots.1067 Figure 3b- represents a zoom in on results after O₂ injection.

Figure 5- Evolution of the active bacterial community (during the incubation at 60 bar, 36°C with a gas mixture (CH₄, 1%CO2, 7.95 ppm of benzene and 3.57 ppm of toluene) before and after oxygen injection (1%). The taxonomic diversity is based on the sequencing of cDNA of the 16S RDNA copies and the quantification was estimated by qPCR of cDNA for each mL of sampled water through the experiment.

1073 Figure 6- X-ray diffraction patterns of samples collected before the experiment (T0) and after
1074 the experiment at three depths (top, middle, bottom). I: illite, K: kaolinite, M: maghemite, Q:
1075 quartz, C: calcite, Ms: marcassite, P: pyrite.

Figure 7- SEM-EDX observations from surface samples collected at the end of experiment
showing clay minerals and quartz (A, SE mode)) and observation of iron sulfides (B, BSE
mode).

The different components:

- 1- Liquid Input
- 2- Gas Output
- 3- Gas temperature probe
- 4- Gas Mixture Input
- 5- O₂ Input
- 6- Stirrer Motor
- 7- Barometer
- 8- Flowmeter
- 9- Barometer
- 10- Gas Output
- 11- O_2 sensor
- 12- Regulation temperature
- 13- Pressure regulation pump
- 14- Liquid Temperature probe
- 15- Liquid Phase
- 16- Piston
- 17- Water from the pressure regulation pump
- 18- Basket containing reservoir rock and capillaries



 Formation water sampling from the well und pressure. Liquid phase preparatiand sampling for microbianalysis and ior chromatography. Solid treatment, capilla preparation, X-r tomography, XRD-SEM. 	 Iger Ionic and gas chromatorafter each liquid and gas Benzene and toluene after each liquid and gas Benzene and toluene after each liquid and gas Liquid sampling for minimized sampling	ngraphy (weekly, before and injection). quantifications (before and injection). crobial analysis (every 10 to after each liquid and gas	 Liquid sampling for microbial analysis and cell culture Solid sampling from the basket for X-ray tomography and XRD- SEM.
 Solid by the react Reactor Liquid p Gas benzender 	t ₀ asket deposition in tor. sterilization. bhase injection. phase (CH ₄ -CO ₂ - c-tolune injection.	 Liquid phase Gas phase ex Solid basket anaerobic co 	e total sampling. vacuation. t removal under inditions.







b)



16S rRNA transcripts

Other

Unknown family

- Thermodesulfovibrionaceae
- Thermoanaerobacteraceae
- Syntrophomonadaceae
- Syntrophaceae
- Staphylococcaceae
- SR-FRB-L83
- Spirochaetaceae
- Sphingomonadaceae
- Solibacteraceae (Subgroup3)
- Ruminococcaceae
- Rhodocyclaceae
- Rhizobiaceae
- Pleomorphomonadaceae
- Peptococcaceae
- Geobacteraceae
- Desulfovibrionaceae
- Corynebacteriaceae
- Bukholderiaceae
- Acetobacteraceae

Time (days)







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Table 1 – Physico-chemical parameters and compounds of the formation water sampled from the deep aquifer (Ab_L_1 site), analyzed at atmospheric pressure.

Dhysica shamical				
Physico-chemical	Value	T T : 4		
Parameters	value	Unit		
Pressure	60	Bar		
Temperature	36	°C		
рН	7.6			
Redox potential	-259	mV		
Conductivity at 25 °C	330	µS_cm		
Organic carbon	0.38	mg.L ⁻¹		
Compounds				
Chloride	0.2475	mM		
Fluoride	0.0043	mM		
Nitrate	0.0034	mM		
Nitrite	0.0005	mM		
Phosphate	0.0041	mM		
Sulfate	0.0835	mM		
Carbonate	0.0083	mM		
Bicarbonate	3.0116	mM		
Calcium	1.1884	mM		
Iron	6.6122	μΜ		
Dissolved iron	5.4486	μΜ		
Magnesium	0.2465	mM		
Potassium	0.1556	mM		
Sodium	0.4930	mM		

Time (Days)	Standard Deviation (16s rRNA transcripts)
0	1.14E+5
43	5.87E+5
48	5.32E+4
85	4.05E+3
103	2.01E+3
107	3.65E+3
130	6.19E+3
179	1.71E+3

Table 2 – Standard deviation related to the 16S rRNA transcripts.

