

# Uptake, translocation, size characterization and localization of cerium oxide nanoparticles in radish (Raphanus sativus L.)

Justyna Wojcieszek, Javier Jiménez-Lamana, Katarzyna Bierla, Lena Ruzik, Monika Asztemborska, Maciej Jarosz, Joanna Szpunar

### ▶ To cite this version:

Justyna Wojcieszek, Javier Jiménez-Lamana, Katarzyna Bierla, Lena Ruzik, Monika Asztemborska, et al.. Uptake, translocation, size characterization and localization of cerium oxide nanoparticles in radish (Raphanus sativus L.). Science of the Total Environment, 2019, 683, pp.284-292. 10.1016/j.scitotenv.2019.05.265. hal-02148731

## HAL Id: hal-02148731 https://univ-pau.hal.science/hal-02148731

Submitted on 25 Oct 2021

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Distributed under a Creative Commons Attribution - NonCommercial 4.0 International License

Version of Record: https://www.sciencedirect.com/science/article/pii/S004896971932306X Manuscript\_320f38f013e9e0158f425b7db4a707b7

### 1 Uptake, translocation, size characterization and localization of

## 2 cerium oxide nanoparticles in radish (*Raphanus sativus* L.)

3

# 4 Justyna Wojcieszek<sup>a</sup>, Javier Jiménez-Lamana<sup>b,\*</sup>, Katarzyna Bierła<sup>b</sup>, Lena Ruzik<sup>a</sup>,

| 5 | Monika A | Asztembors | ka <sup>c</sup> , N | laciej J | arosz <sup>a</sup> , | Joanna | Szpunar <sup>®</sup> |
|---|----------|------------|---------------------|----------|----------------------|--------|----------------------|
|---|----------|------------|---------------------|----------|----------------------|--------|----------------------|

- <sup>6</sup> <sup>a</sup> Chair of Analytical Chemistry, Faculty of Chemistry, Warsaw University of Technology, Poland
- 7 <sup>b</sup> Institute of Analytical Sciences and Physico-Chemistry for Environment and Materials (IPREM),
- 8 CNRS-UPPA, UMR5254, Pau, France
- <sup>9</sup> <sup>c</sup> Isotopic Laboratory, Faculty of Biology, University of Warsaw, Warsaw, Poland
- 10 \*Corresponding author: j.jimenez-lamana@univ-pau.fr Telephone: +33540175037
- 11

#### 12 Abstract

Due to their unique physical and chemical properties, the production and use of cerium oxide 13 14 nanoparticles (CeO<sub>2</sub> NPs) in different areas, especially in automotive industry, is rapidly increasing, causing their presence in the environment. Released CeO<sub>2</sub> NPs can undergo different transformations 15 16 and interact with the soil and hence with plants, providing a potential pathway for human exposure and leading to serious concerns about their impact on the ecosystem and human organism. This study 17 18 investigates the uptake, bioaccumulation, possible translocation and localization of CeO2 NPs in a model plant (Raphanus sativus L.), whose edible part is in direct contact with the soil where 19 20 contamination is more likely to happen. The stability of  $CeO_2$  NPs in plant growth medium as well as after applying a standard enzymatic digestion procedure was tested by single particle ICP-MS (SP-21 ICP-MS) showing that CeO<sub>2</sub> NPs can remain intact after enzymatic digestion; however, an 22 agglomeration process was observed in the growth medium already after one day of cultivation. An 23 enzymatic digestion method was next used in order to extract intact nanoparticles from the tissues of 24 25 plants cultivated from the stage of seeds, followed by size characterization by SP-ICP-MS. The results obtained by SP-ICP-MS showed a narrower size distribution in the case of roots suggesting preferential uptake of smaller nanoparticles which led to the conclusion that plants do not take up the CeO<sub>2</sub> NPs agglomerates present in the medium. However, nanoparticles at higher diameters were observed after analysis of leaves plus stems. Additionally, a small degree of dissolution was observed in the case of roots. Finally, after CeO<sub>2</sub> NPs treatment of adult plants, the spatial distribution of intact CeO<sub>2</sub> NPs in the radish roots was studied by laser ablation ICP-MS (LA-ICP-MS) and the ability of NPs to enter and be accumulated in root tissues was confirmed.

33

Keywords: Single particle ICP-MS, edible plants, engineered nanoparticles, laser ablation ICP-MS,
ionization efficiency

36

#### 37 **1. Introduction**

In recent years there has been a rapid increase in the development, production and application of engineered nanoparticles (ENPs), which are currently playing a growing role in a wide range of industrial and commercial applications (Stark et al., 2015; Vance et al., 2015). Due to their unique properties and novel features, the global production of nanomaterials is significantly increasing, leading to the release of ENPs into the environment, where they can interact with higher plants.

Among the different ENPs, cerium oxide nanoparticles (CeO<sub>2</sub> NPs) are one of the most 43 important and promising due to their capability for undergoing a redox-cycle between the two natural 44 oxidation states (Ce<sup>3+</sup> and Ce<sup>4+</sup>) (Cassee et al., 2011), with an estimated global production of up to 10 45 46 000 tons per year (Keller et al., 2013; Piccinno et al., 2012). Cerium is the most abundant rare earth 47 element in the earth's crust and belongs to the group of elements referred to as technology-critical elements that were considered just as laboratory curiosities, being now key components for the 48 49 development of new technologies (Wojcieszek et al., 2018). CeO<sub>2</sub> NPs have special electrical, optical 50 and thermal properties which make them widely used as polishing agents, UV-blockers, glass additives, in agricultural products and especially in automotive industry (Stanek et al., 2008; The 51 Project on EmergingNanotechnologies, 2019; Yang et al., 2017). Moreover, the high reactive oxygen 52

53 species (ROS) scavenging properties gives  $CeO_2$  NPs the potential to be used as an antioxidant and radioprotective agent (Li et al., 2015). For instance, CeO<sub>2</sub> NPs are widely used in the production of 54 catalysts or as a diesel fuel additive to increase fuel combustion efficiency, decrease diesel soot 55 emissions and to reduce NOx emissions (Cassee et al., 2011; Johnson and Park, 2012). Therefore, 56 57 their release into the ecosystem is unavoidable for example through deposition along roadways that leads to their translocation to the soil and thus to plants, which has raised serious concerns about their 58 59 fate and impact in the ecosystem (Colvin, 2003; Gardea-torresdey et al., 2014; Johnson and Park, 60 2012). CeO<sub>2</sub> NPs can interact with the soil and hence with plants, which in turn increase the risk of 61 their bioaccumulation in the animal and human food chain (Colvin, 2003; Deng et al., 2014). As a 62 results, and due to the importance of the physicochemical states of Ce to its food safety implications, a 63 more extensive studies on the interaction of  $CeO_2$  NPs and higher plants is essential for a fully 64 understanding of their environmental impact (W. Zhang et al., 2017a).

65 Both positive and negative effects of CeO<sub>2</sub> NPs on plants have been demonstrated, depending 66 on the plant species, size and concentration of NPs, exposure time and plant growth conditions (Ma et 67 al., 2010; Rossi et al., 2016; Wang et al., 2012; Zhao et al., 2014). Species-dependent responses of 68 plants for CeO<sub>2</sub> NPs treatment have been also observed by other researchers (Lopez-Moreno et al., 69 2010b). Additionally, it was already presented that toxicity of  $CeO_2$  NPs towards plants is 70 concentration dependent (Lopez-Moreno et al., 2010a). On the other hand, CeO<sub>2</sub> NPs were shown to 71 enhance plant growth under certain exposure conditions (Ma et al., 2016; Wang et al., 2012). CeO<sub>2</sub> 72 NPs have generally been considered as insoluble under environmental conditions (Walser et al., 2012) 73 (Yokel et al., 2009). However, several studies suggest that after CeO<sub>2</sub> NPs treatment, cerium in plants 74 could exist in the form of Ce salts or dissolved Ce; reducing agents present in plant growth media could reduce the surface  $Ce^{4+}$  of  $CeO_2$  NPs into  $Ce^{3+}$  (Dan et al., 2016; Schwabe et al., 2014; Zhang et 75 al., 2012; W. Zhang et al., 2017). Additionally, solubility of CeO<sub>2</sub> NPs is also pH dependent 76 77 (Hernandez-viezcas et al., 2013). Dissolved form of CeO<sub>2</sub> NPs was found for example in cucumber or soybean plants (Hernandez-Viezcas et al., 2016; Zhang et al., 2012). In contrast, other studies showed 78 that CeO<sub>2</sub> NPs did not undergo any transformation in cucumber, alfalfa, tomato and corn seedling 79 80 (Lopez-Moreno et al., 2010b). Most of the studies performed showed that CeO<sub>2</sub> NPs can be taken up by plants, but the majority of NPs appeared to remain in root tissues, raising concerns on their
accumulation by root vegetables. However, even though the edible tissues of below-ground
vegetables have direct contact with NPs contaminated soils, only little attention has been paid to this
group of plants (Zhang et al., 2015).

85 In this context, radish (Raphanus sativus L.), a popular vegetable with high global consumption, was chosen to the study since its edible part is in contact with the soil where 86 contamination, through atmospheric dust deposition, is more likely to happen. Taking into account 87 different, often contradictory, results presented in previous works, the main purpose of this study was 88 to investigate the uptake, translocation, characterization and spatial localization of a commercial 89 suspension of  $CeO_2$  NPs in a model edible plant throughout the whole process by using different mass 90 spectrometry based techniques: the stability of the  $CeO_2$  NP in the hydroponic solution was assessed 91 92 by single particle ICP-MS (SP-ICP-MS); the bioaccumulation of cerium in plants tissues and the translocation factor from roots to leaves was determined by ICP-MS; the presence of  $CeO_2$  NPs in 93 radish tissues was investigated by SP-ICP-MS and their size distributions obtained; finally, the spatial 94 95 distribution of cerium in roots was studied by laser ablation ICP-MS (LA-ICP-MS).

96

#### 97 2. Material and methods

#### 98 2.1. Samples and reagents

99 Seeds of radish (Raphanus sativus L.) were purchased from Vilmorin Garden (Komorniki, 100 Poland). Analytical or biological reagent grade chemicals and LC-MS grade solvents were purchased 101 from Sigma-Aldrich (Saint Quentin Fallavier, France) unless stated otherwise. Ultra-pure water (18 MΩ cm) obtained with a MiliQ system (Millipore, Guyancourt, France) was used 102 103 throughout. Hydrogen peroxide from Fisher Scientific (Hampton, NH) and nitric acid (INSTRA-Analysed) from Baker (Deventer, Netherlands) were used for samples digestion. A standard ionic 104 solution of 1000 mg L<sup>-1</sup> cerium was purchased from Plasma CAL standards (Teddington, UK). 105 Macerozyme R-10 enzyme (pectinase from Rhizopus sp., Sigma Aldrich) was used to digest plant 106

tissues for CeO<sub>2</sub> NPs extraction. Macerozyme R-10 is a multi-component enzyme mixture containing
cellulase (0.1 unit per mg), hemicellulase (0.25 unit per mg) and pectinase (0.5 unit per mg).

Aqueous suspension of CeO<sub>2</sub> NPs (40 wt %) with a nominal size of 30-50 nm was purchased from US Research Nanomaterials, Inc. (Houston, TX). A diluted suspension of CeO<sub>2</sub> NPs was prepared daily in ultrapure water by accurately weighing aliquots of the stock suspension after one minute sonication. After dilution and before each analysis, the suspensions were sonicated for approximately 1 min.

#### 114 2.2. Instrumentation

115 Homogenization of the plant tissues was performed by a Vibracell 75115 ultrasonic probe (Bioblock Scientific, Illkirch, France) offering a nominal power of 500 W. Incubation of samples was 116 performed in a Grant OLS-200 water bath (Keison Products, Essex, UK). A Branson B2510 ultrasonic 117 bath (Branson, Danbury, CT) was used for sonication of nanoparticles suspensions and enzymatic 118 119 extracts before SP-ICP-MS analysis. A Digi-Prep system from SCP Science (Quebec, Canada) was used for acid digestion of plant tissues and extracts for the total cerium determination. A centrifuge 120 5415R (Eppendorf, Hamburg, Germany) was used for the extraction procedure. pH was adjusted by 121 122 using a FiveEasy pH meter from Mettler Toledo (Columbus, OH).

123 The determination of the total concentration of cerium in the plant samples was carried out 124 using an Agilent 7500ce ICP-MS (Agilent, Tokyo, Japan) instrument fitted with Ni cones and a 2.5 125 mm i.d. injector torch. The position of torch and nebulizer gas flow was adjusted each day of work. 126 The working conditions of ICP MS were optimized daily using a 1  $\mu$ g L<sup>-1</sup> solution of <sup>7</sup>Li<sup>+</sup>, <sup>89</sup>Y<sup>+</sup> and 127 <sup>205</sup>Tl<sup>+</sup> in 2% (v/v) HNO<sub>3</sub>.

128 2.3. Single particle ICP-MS method

An Agilent 7900 ICP MS equipped with Single Nanoparticle Application Module was used for the characterization of CeO<sub>2</sub> NPs. The default instrumental and data acquisition parameters are listed in Table 1. During analysis <sup>140</sup>Ce with a natural abundance of about 88.5 % was monitored. <sup>140</sup>Ce is interference-free in ICP-MS analysis which makes this isotope the best choice for analysis of Ce based nanoparticles. SP-ICP-MS analyses were performed in TRA mode using a dwell time of 100 134  $\mu$ s, with a total time of analysis of 60 s. The dwell time is a critical parameter when working in single 135 particle mode and the use of microsecond dwell time can improve particle sizing accuracy and counting (Abad-Álvaro et al., 2016). Gold nanoparticle standard reference material with a nominal 136 137 diameter of 56 nm (RM 8013) was obtained from NIST (Gaithersburg, USA) and was used for determination of transport efficiency, which was calculated by the particle frequency method 138 described by Pace et al. (2011). The sample flow rate was calculated daily by measuring the mass of 139 water taken up by the peristaltic pump for 2 min (this operation was repeated 3 times). After each 140 sample analysis, the software automatically processed the raw data and generated the particle size, 141 particle concentration, size distribution and information about concentration of dissolved metal. The 142 size distributions were prepared in Origin 8.5 software (Northampton, MA) and adjusted to lognormal 143 distributions in order to obtain the median diameter. 144

145

#### 146 Table 1

| 147 Default instrumental and data acquisition parameters for SP-ICP-I |
|---|
|---|

| Instrumental parameters     |                           |  |
|-----------------------------|---------------------------|--|
| RF Power                    | 1550 W                    |  |
| Argon gas flow rate         |                           |  |
| Plasma                      | 15 L min <sup>-1</sup>    |  |
| Auxiliary                   | 0.9 L min <sup>-1</sup>   |  |
| Nebulizer                   | 1.10 L min <sup>-1</sup>  |  |
| Sample uptake rate          | 0.35 mL min <sup>-1</sup> |  |
| Data acquisition parameters |                           |  |
| Dwell time                  | 100 µs                    |  |
| Readings per replicate      | 600000                    |  |
| Total acquisition time      | 60 s                      |  |
| Analyte                     | Ce                        |  |
| Mass (amu)                  | 140                       |  |
| Density                     | 7.13 g cm <sup>-3</sup>   |  |
| Mass fraction               | 0.8139                    |  |

148 149

150 2.4. Laser ablation ICP-MS method

151 A NewWave UP-213 laser ablation (LA) system (NewWave Research, Freemont, USA) was coupled to an Agilent model 7700x (Agilent, Tokyo, Japan) using a 60 cm Tygon tube (5.0 mm i.d.). 152 The laser ablation system was operated in a focused spot mode at the repetition rate of 20 Hz with a 153 spot size of 250  $\mu$ m and a scan speed of 100  $\mu$ m s<sup>-1</sup>. The ablated matter was transported into the ICP 154 155 with He gas (800 mL min<sup>-1</sup>) and mixed in a T-connector with carrier gas of ICP MS delivered at 1.1 L min<sup>-1</sup> aerosol (a Micromist nebulizer and a double pass Scott spray chamber were removed from the 156 configuration). No collision/reaction gas was used during analyses. A 1.5 mm i.d. injector torch and 157 Pt cones were used. Pulse energy was of 30% and fluence was 1.45 J cm<sup>-2</sup>. The ICP-MS instrument 158 159 was optimized in wet plasma condition before coupling with laser in order to obtain the maximum Ce 160 signal-to-noise ratio.

161 *2.5. Procedures* 

#### 162 2.5.1. Plant cultivation

Portions of radish seeds (ca. 1 g) were germinated in distilled water in darkness for 3 days and 163 then the seedlings were transferred to 350 mL opaque containers with a Knop nutrient solution (pH 164 6.8) and placed in a growth chamber. After 4 days, the CeO<sub>2</sub> NPs suspension was added to the 165 medium at a cerium concentration of 5 mg  $L^{-1}$ . As a control variant, plants were left in the nutrient 166 solution without the addition of cerium. Cultivation was carried out for the next 7 days in a growth 167 chamber. The plants were grown at with a temperature varying from 21-25 °C, light intensity 168 169 200 microeinsteins m<sup>-2</sup> s<sup>-1</sup>, 12 h photoperiod and 50-60% relative humidity. All the time the solutions 170 in the containers were aerated. Each variant of the cultivation was performed in three replicates. After cultivation, the plants were harvested and the roots were gently rinsed with deionized water. The 171 172 plants were divided into roots and above ground organs (leaves plus stems) and lyophilized. Dried plant material was ground in a mortar before further analysis. 173

For LA-ICP-MS analysis, radish plants were initially cultivated in typical garden soil in the greenhouse for 4 weeks to obtain well-formed tubers. Next, adult plants of radish, with the taproot diameter of ca. 1.5 cm were conditioned in deionized water for 24 hours and then transferred to opaque containers with suspensions of CeO<sub>2</sub> NPs (50 mL) at concentrations of 5 mg L<sup>-1</sup> or 50 mg L<sup>-1</sup> for 24 and 48 h. After finishing the cultivation, taproots were gently rinsed with deionized water andprepared for further analysis. Each variant of the cultivation was performed in three replicates.

#### 180 2.5.2. Determination of total cerium content

181 Samples of leaves and roots (0.025 g) were digested by adding 2.5 mL HNO<sub>3</sub> (c) in a DigiPrep tube and the following temperature program: 30 min of heating up to 65°C and then keeping 182 183 temperature at  $65^{\circ}$ C for 4 h. Afterwards, 1 mL of H<sub>2</sub>O<sub>2</sub> (c) was added to the samples and the digestion was continued for the next 4 h using the same program. After digestion, the mixtures were cooling 184 down and diluted with ultrapure water to a final volume of 25 mL. Further dilutions were prepared 185 with 2% HNO<sub>3</sub> (v/v), directly before ICP-MS analysis. Two isotopes of cerium (<sup>140</sup>Ce, <sup>142</sup>Ce) were 186 monitored during analysis. The analytical blanks were analyzed in parallel. Quantification was 187 188 performed by external calibration; a 5 points calibration curve was prepared for cerium concentration in the investigated range from 0.0 to 5.0 ng mL<sup>-1</sup>. The content of cerium in each sample was 189 190 calculated as the mean of results obtained for the monitored isotopes, from three replicates of each sample. In order to verify the accuracy of the method, a recovery study was performed. A known 191 192 amount of CeO<sub>2</sub> NPs suspension (at the same concentration as used for plant cultivation) was spiked 193 into the controls of roots and leaves plus stems and the acid digestion method described above was 194 applied.

#### 195 2.5.3. Enzymatic digestion method

After cultivation performed on radish seeds, plant tissues were digested enzymatically as 196 197 reported in previous works (Jiménez-Lamana et al., 2016; Kińska et al., 2018). Briefly, grounded samples of leaves and roots (0.025 g) were homogenized with 8 mL of 2 mM citrate buffer (pH 4.5; 198 199 adjusted with citric acid) by using an ultrasonic probe. After the end of homogenization, 2 mL of 200 Macerozyme R-10 solution (0.01 g of enzyme powder for roots and 0.05 g of enzyme powder for 201 leaves plus stems, dissolved in 2 mL of ultrapure water) was added to samples and they were next 202 shaken at 37 °C for 24 hours in a water bath with continuous shaking. After digestion, the samples were settle down for approximately 15 min and the obtained suspensions were filtered with a 0.45 µm 203 syringe filter (Sigma Aldrich). The filtered samples were next analyzed by SP-ICP-MS. 204

205 2.5.4. Study of the spatial distribution of CeO<sub>2</sub> NPs in radish roots by LA-ICP-MS

| 206 | After the end of cultivation performed on adult plants, radish roots were washed with                 |
|-----|---|
| 207 | ultrapure water in order to remove nanoparticles adsorbed on the surface. Afterwards, fresh roots     |
| 208 | samples were cut into 150 $\mu$ m thick slices using the Leica VT1000 S vibrating blade microtome     |
| 209 | (Leica Biosystems, Wetzlar, Germany). The radish slices were directly mounted onto glass slides       |
| 210 | coated with double-sided tape to enhance adhesion and kept in a fridge before further treatment. Four |
| 211 | repetitions of each root sample were prepared. The prepared slices were next analyzed by means of     |
| 212 | LA-ICP-MS. The tissue section was systematically scanned (line by line) by a focused laser beam.      |

213

#### 214 **3. Results and discussion**

# 3.1. Cultivation of radish in the presence of nanoparticles and determination of total content of cerium

217 During cultivation, radish plants appeared tolerant to the applied concentration of  $CeO_2$  NPs 218 since no visible phytotoxic effects were observed. Biomass production and tissues hydration did not 219 differ substantially between the control and the  $CeO_2$  NPs treated plants. Some additional studies on 220 the content of photosynthetic pigments showed that plants were not affected by nanoparticles.

The determination of the total cerium content in radish tissues was the first step done after the plant cultivation with the CeO<sub>2</sub> NPs suspension, in order to investigate its distribution among the different tissues. Control and CeO<sub>2</sub> NPs treated samples of leaves (together with stems) and roots of radish were analyzed by standalone ICP-MS after samples mineralization. The quantification limit, calculated as 10 times the standard deviation of the blank (n = 3) divided by the sensitivity, was 0.43 ng g<sup>-1</sup>. The results of cerium content were obtained from three independent experiments and are presented in Table 2.

228

#### 229 Table 2.

Total cerium content in radish tissues from control and  $CeO_2$  NPs treated plants (mean  $\pm$  standard deviation).

|                | Control plant samples /µg g-1 | CeO <sub>2</sub> NPs treated samples /µg g <sup>-1</sup> |
|----------------|-------------------------------|--|
| Leaves + stems | < LOQ                         | $32.9 \pm 0.4$   |
| Roots          | < LOQ                         | $1948 \pm 9$   |

232

233 The obtained results showed that cerium is accumulated in analyzed radish tissues. However, 234 the majority of cerium was found in radish roots and only minor part of cerium was transported to 235 above ground organs, which is in good agreement with many studies as discussed below. The translocation factor (TF), defined as the ratio of cerium content in leaves with respect to roots, was 236 237 0.017. It has been reported that the translocation from roots to shoots is independent of the cerium 238 content in the roots and of the supplied form of cerium (Yang et al., 2017). Taking into account that 239 roots as edible part of radish are in direct contact with the soil where contamination is more likely to 240 happen, the results of the total content are especially important from the point of view of food safety. 241 To determine the accuracy of the proposed acid digestion method, a recovery study was made. 242 Control plant samples (leaves plus stems, and roots) were spiked with a known amount of CeO<sub>2</sub> NPs and the acid digestion method was applied. Good recovery values were obtained for both leaves plus 243 stems (103.2%) and roots (102.7%). Regardless of the total content of cerium determined in plant 244 245 tissues in previous works, its translocation from roots to the aerial part was always in very low level. 246 For instance, Yang et al. (2017) reported that the TF of cerium from roots to above ground organs in Arabidopsis thaliana plant is 0.042, while Rossi et al. (2016) found a TF of around 0.002 after 247 analysis of Brassica napus L. In the case of desert plant mesquite, the TF of cerium was 0.2 and 0.25 248 for plants treated with 500 and 4000 mg L<sup>-1</sup>, respectively (Hernandez-Viezcas et al., 2016). In the case 249 250 of plants grown in soils, a TF of 0.02–0.03 was found in kidney bean plants (Majumdar et al., 2016). It should be also mentioned that no translocation of cerium to above ground organs was observed in 251 the case of tomato or wheat plants (Antisari et al., 2015; Schwabe et al., 2015). The TF is dependent 252 also on the size of  $CeO_2$  NPs. It was showed that after analysis of pumpkin leaves and roots, the TF 253 254 for CeO<sub>2</sub> NPs of 9 nm was 0.0004 whereas no cerium translocation was observed in the case of NPs of 255 64 nm (Schwabe et al., 2015).

256 3.2. Study of ionization efficiency of CeO<sub>2</sub> NPs in the plasma

257 The calculations behind SP-ICP-MS theory lie on some assumptions, being a plasma ionization efficiency for nanoparticles comparable to the corresponding dissolved species one of them 258 (Pace et al., 2011). However, while this has proven to be true for silver or gold nanoparticles (Hu et 259 al., 2009; Laborda et al., 2011), it might not be the case for nanoparticles of a different nature, like it 260 261 has been reported for selenium nanoparticles (Jiménez-Lamana et al., 2018), and needs to be first investigated. A ionization efficiency lower than 100% was reported for CeO<sub>2</sub> NPs with sizes from 1 to 262 10 nm (Sánchez-garcía et al., 2016). However, this value might be different in the case of the  $CeO_2$ 263 NPs used in this study, since the ablation and ionization may strongly vary while increasing the size 264 of the nanoparticle. In order to calculate the ionization efficiency, the response of the ICP-MS towards 265 the nanoparticulated form of cerium was studied, by determining and comparing the total cerium 266 concentration of digested nanoparticles and a suspension of undigested nanoparticles. The calibration 267 268 was achieved with aqueous standards of cerium in 2% HNO<sub>3</sub> for the digested nanoparticles and in ultrapure water and in 2% HNO<sub>3</sub> for the direct analysis of the suspension, in order to investigate the 269 270 influence of the medium on the cerium sensitivity. The concentrations determined in ultrapure water 271 and in 2% HNO<sub>3</sub> for the undigested CeO<sub>2</sub> NPs suspension were 9.7  $\pm$  0.9% and 10.2  $\pm$  1.2% with 272 regard to the concentration determined after acid digestion, respectively (Table 3).

- 273
- 274 **Table 3.**

<sup>275</sup> Determination of cerium concentration in digested and undigested suspension of  $CeO_2$  NPs (mean  $\pm$ 276 standard deviation).

| Sample                                   | Total Cerium content / mg L <sup>-1</sup> |  |
|--|---|--|
| Acid digestion                           | $237.9 \pm 3.0$                           |  |
| Suspension in ultrapure water            | $23.1 \pm 2.1$                            |  |
| Suspension in 2 % HNO <sub>3</sub> (v/v) | $24.4 \pm 2.9$                            |  |

<sup>277</sup> 

These results show that ICP-MS sensitivity depends on the physicochemical form of cerium and thus a particle ionization efficiency, defined as the ratio of the ionization efficiency of the particle to the ionization efficiency of the corresponding dissolved metal solution (Pace et al., 2011), needs to be taken into account. According to the results obtained in this study, a particle ionization efficiency of 10% was applied for sizing CeO<sub>2</sub> NPs. In addition, it was shown that the medium has no significant influence on the ionization of CeO<sub>2</sub> NPs.

The suspension of  $CeO_2$  NPs used in this study was then analysed by means of SP-ICP-MS at a concentration of around  $1x10^8$  NP L<sup>-1</sup> taking into account the ionization efficiency calculated. The size distribution obtained (Fig. 1) showed a polydisperse distribution with its central part within the size range provided by the manufacturer: 30-50 nm.



288

Fig. 1. Size distribution of the commercial suspension of CeO<sub>2</sub> NPs used in the cultivation studies.

- 291 *3.3 Investigation of CeO*<sub>2</sub> *NPs stability*
- 292 *3.3.1 Stability in growth medium*

Before the analysis of plant samples, the stability of the suspension of  $CeO_2$  NPs in the 293 growth medium was investigated. A suspension of CeO<sub>2</sub> NPs of 5 mg L<sup>-1</sup> was spiked into the nutrient 294 solution and the size distribution of the nanoparticles was determined by SP-ICP-MS immediately 295 upon preparation and after 1, 2, 4 and 7 days. Each suspension analysed was diluted accordingly in 296 297 order to measure the same number of events. The median diameters obtained from the corresponding size distributions of nanoparticles at the different times analysed are shown in Table 4. From the 298 results obtained, it can be concluded that  $CeO_2$  NPs undergo agglomeration after 1 day of cultivation 299 in growth medium, with an increase of the median diameter from 56.0 nm to 128.3 nm, whereas the 300

301 size distribution does not vary significantly from 1 to 7 days of cultivation (Fig. S1). The agglomeration of CeO<sub>2</sub> NPs has already been observed in hydroponic solution, without the addition of 302 a stabilizing agent (Schwabe et al., 2013; Schwabe et al., 2014). Remarkably, no background signal 303 corresponding to dissolved cerium was observed on the SP-ICP-MS time scans at any of the 304 305 incubation times tested in the present study. This finding can be explained by two facts: on one hand, the occurrence of CeO<sub>2</sub> NPs agglomerates prevent from cerium dissolution, since agglomeration 306 reduces the effective specific surface area of nanoparticles, which results in less cerium dissolution; 307 on the other hand, the release of cerium ions is size dependent: in a study performed with  $CeO_2$  NPs 308 in Hoagland medium, Schwabe et al. (2014) showed that the smallest particles (9 nm) released the 309 310 highest amount of cerium, whereas the release of cerium by the largest particles (64 nm) was almost 311 negligible.

312

#### 313 Table 4.

314 Median diameters of CeO<sub>2</sub> NPs suspended in nutrient solution after different contact times.

| Incubation time | Median diameter / nm |
|-----------------|----------------------|
| 0               | $56.0 \pm 1.9$       |
| 1 day           | $128.3 \pm 19.1$     |
| 2 days          | $119.1 \pm 4.6$      |
| 4 days          | $112.5 \pm 3.9$      |
| 7 days          | $134.6 \pm 22.7$     |

315

316 *3.3.2. Influence of enzymatic digestion and plant matrix* 

The influence of the enzymatic procedure on the size distribution of the CeO<sub>2</sub> NPs suspension was also studied. For this purpose, a suspension of CeO<sub>2</sub> NPs (5 mg  $L^{-1}$ ) with no plant tissue was treated with the same conditions of enzymatic digestion. As it can be shown in Fig. 2, the size distribution obtained for enzyme-treated CeO<sub>2</sub> NPs was in good agreement with the size distribution obtained for CeO<sub>2</sub> NPs stock suspension freshly prepared and hence the integrity of CeO<sub>2</sub> NPs is not affected by the enzyme used in the digestion procedure.

In addition, the influence of the plant matrix was investigated. Control leaves plus stems and control roots of radish were spiked with 5 mg  $L^{-1}$  of the CeO<sub>2</sub> NPs suspension, submitted to the enzymatic digestion procedure and the corresponding nanoparticle size distributions obtained by SPICP-MS. The size distributions obtained for both spiked tissues were in good agreement with those
obtained for freshly enzyme-treated CeO<sub>2</sub> NPs (Fig. 2). These results showed that the procedure
proposed can be used to extract CeO<sub>2</sub> NPs from plant tissues without causing their transformation.



330

**Fig. 2.** Size distributions of: CeO<sub>2</sub> NPs stock suspension freshly prepared (blue); CeO<sub>2</sub> NPs after enzymatic digestion procedure (orange); control leaves plus stems (green) and roots (yellow) spiked with 5 mg  $L^{-1}$  CeO<sub>2</sub> NPs suspension. Size distributions were adjusted to lognormal distributions (grey line) and the median diameter ± the standard deviation obtained.

335

#### 336 *3.4. Single particle ICP-MS analysis of leaves plus stems and roots of radish*

In order to identify the form of cerium inside the radish tissues cultivated from the stage of seeds, samples of leaves plus stems and samples of roots were subjected to the enzymatic procedure followed by SP-ICP-MS analysis. In a first step, roots and leaves plus stems of control plants were analyzed and the corresponding time scans obtained (Fig. 3a and 3b, respectively). As expected, only
a few pulses above the background were observed for control plants. The presence of these pulses
may be explained by a slight contamination of the sonication probe during the enzymatic digestion
procedure, but their occurrence is not significant (less than 10 pulses out of 600,000 readings) and
hence does not hamper the further analyses.

345

346



Fig. 3. Time scans obtained for samples of: a) roots, and b) leaves of control plants and for c) roots
and d) leaves of plants treated with 5 mg L<sup>-1</sup> of CeO<sub>2</sub> NPs.

Afterwards, the SP-ICP-MS analyses of CeO<sub>2</sub> NPs treated samples were performed. Samples 350 351 of roots were diluted 3000 times before analysis whereas samples of leaves plus stems were diluted 352 200 times. Each tissue was analyzed by triplicate. Time scans obtained showed a significant number of pulses in both plants tissues (Fig. 3c and 3d), proving that radish is able to uptake CeO<sub>2</sub> NPs and 353 transports them into the aerial part of the plant. The presence of  $CeO_2$  NPs has been already reported 354 in tomato (S. lycopersicum L.), cucumber (C. sativus), pumpkin (Cucurbita pepo), and soybean 355 (Glycine max) shoots (Dan et al., 2016; Zhang et al., 2011; Zhang et al., 2017b), in Arabidopsis 356 thaliana shoots (Yang et al., 2017) and in romain lettuce shoots (P. Zhang et al., 2017a). According to 357

358 the number of pulses detected, the nanoparticle number concentration in roots and in leaves plus stems was calculated as 3.3 x 10<sup>11</sup> and 2.4 x 10<sup>10</sup> NP L<sup>-1</sup>, respectively. The largest number of number 359 concentration of nanoparticles observed in roots could be explained by a direct contact of roots with 360  $CeO_2$  NPs suspension. Remarkably, the intensity of the pulses observed in the time scan obtained for 361 362 leaves is significantly higher than for roots, which suggested the presence of bigger nanoparticles. In addition, no signal corresponding to dissolved cerium was observed for leaves, suggesting that  $CeO_2$ 363 NPs did not underwent dissolution during their transport, although small signal coming from 364 365 dissolved form of cerium was observed for roots. This is in good agreement with other studies that have shown that the dissolution of  $CeO_2$  NPs occur at the root surface rather than inside plants (P. 366 367 Zhang et al., 2017b). For instance, Ma et al. (2015) shown that dissolution of  $CeO_2$  NPs only occurred at root surfaces, whereas Ce (IV) was not reduced in the tissues in hydroponic cucumber plants. Other 368 369 authors have reported that plants roots activity have an impact on the dissolution of  $CeO_2$  in hydroponically grown wheat, pumpkin, and sunflower plants (Schwabe et al., 2015). 370

371 Due to the small signal coming from dissolved form of cerium observed after analysis of roots 372 by SP-ICP-MS, a speciation study was performed by means of size exclusion chromatography (SEC) 373 coupled to ICP-MS. The chromatogram obtained (Fig. S1) showed just one signal coming from high 374 molecular weight cerium compounds which could be explained by the ability of metal ions to create 375 agglomerates even with low molecular weight ligands. However, the extraction efficiency of cerium 376 from radish roots (procedure explained in Supplementary Information), was determined at only 0.19% 377 and therefore, due to negligible presence of cerium compounds, the speciation study of cerium was 378 not continued.

From time scans and by applying the ionization efficiency factor calculated before, the size distributions of  $CeO_2$  NPs in radish tissues were obtained. As it can be observed, the particle size distribution obtained for radish roots was narrower (Fig. 4a), with a median diameter of 42.7 nm, suggesting a preferential uptake of NPs at smaller sizes and that the agglomerates present in growth medium already after one day are not taken up by the plant. The finding that plant uptake of  $CeO_2$ NPs depends on the particle size and smaller particles are more readily taken up by plants has been previously suggested (Zhang et al., 2011). However, the particle size distribution obtained for leaves 386 plus stems (Fig. 4b), showed a significant number of nanoparticles at higher diameters than in the case 387 of roots (median diameter: 60.3 nm). Since it was found that plants are not able to take up big nanoparticles this finding may be explained by an agglomeration process that takes place at the 388 389 endpoint of transport of CeO<sub>2</sub> NPs from roots to above ground organs. Once CeO<sub>2</sub> NPs have entered 390 the plant, they are translocated up to above-ground organs through the xylem system, driven by the 391 transpiration stream (Zhao et al., 2013). The endpoint of this transport pathway is the leaves (Zhao et al., 2013), where nutrients and water arrive through the veins. At this point, nanoparticles will be 392 locally more concentrated and hence agglomeration by contact is more likely to occur. Similarly, 393 particle aggregates have been found in leaves of Arabidopsis thaliana treated with CeO<sub>2</sub> NPs (Yang et 394 al., 2017). 395

396

397



**Fig. 4.** Size distributions obtained for a) roots, and b) leaves plus stems of radish

399

The biotransformation of CeO<sub>2</sub> NPs inside plants remains a controversial issue. Some authors 400 401 have suggested the uptake of Ce(III) followed by re-precipitation as pathway in CeO<sub>2</sub> NPs uptake by 402 plants (Schwabe et al., 2015). However, our study showed that the dissolution of cerium is not significant and thus this mechanism can be discarded. On the other hand, cerium has been shown to be 403 present in the shoots of cucumber and lettuce plants as CeO<sub>2</sub> and cerium carboxylates (Zhang et al., 404 2012; Zhang et al., 2017a), although the majority of studies have shown that cerium mostly remained 405 as CeO<sub>2</sub> NP in plants tissues (Majumdar et al., 2014; Zhang et al., 2017b). For instance, in a study 406 conducted with romaine lettuce, the amount of cerium carboxylates in leaves was reported at 3.5% (P. 407 408 Zhang et al., 2017a). These previous findings make us state that the cerium signal observed in radish 409 leaves plus stems is due to CeO<sub>2</sub> NPs and/or aggregates.

#### 410 3.5. Study of spatial distribution of CeO<sub>2</sub> NPs in radish roots by LA-ICP-MS

The results obtained by SP-ICP-MS analysis showed that the great majority of CeO<sub>2</sub> NPs 411 remain in the form of nanoparticles both in roots and leaves. Therefore, in order to check if CeO<sub>2</sub> NPs 412 413 have the ability to enter and be accumulated inside the radish tissues, the localization of  $CeO_2$  NPs in 414 the radish roots, as edible part of the plant, was determined by LA-ICP-MS. Analyses were performed 415 after cultivation of radish carried out on adult plants. It is worth to mention that radish belongs to the 416 group of plants with root system in type of taproot, with one central and dominant root. Additionally, 417 taproot of radish is in the shape of fusiform root – The primary root of the system is the widest in the middle with secondary root tapers towards the bottom. Optimisation of a LA-ICP-MS analysis in 418 terms of sensitivity and S/N ratio allowed to choose optimum scan speed of 100 µm s<sup>-1</sup> and spot size 419 420 of 250 µm; increase in the laser beam spot size resulted in a considerable gain in sensitivity and homogeneity of the signal as more material is introduced into the plasma per time unit. After analysis 421 422 of control plants cultivated for 2 days, no signal corresponding to cerium was observed (Fig. 5). The results obtained for radish samples treated with 5 mg L<sup>-1</sup> of CeO<sub>2</sub> NPs (Fig. 5A) showed peaks from 423 424 cerium only at the external part of the analysed slices. Nanoparticles were found at a depth of about

425 1.2 mm leading to the conclusion that  $CeO_2$  NPs have the ability to enter and be accumulated into the 426 radish tissues, which is in good agreement with another study performed with radish in soil (W. Zhang et al., 2017b). Similar situation was observed in the case of radish samples treated with 50 mg 427  $L^{-1}$  CeO<sub>2</sub> NPs (Fig. 5B). Signals from cerium were registered just below the skin surface, with higher 428 429 intensity than those observed for samples treated with 5 mg  $L^{-1}$  CeO<sub>2</sub> NPs. However, one additional signal was observed in the central part of samples after 1 day and 2 days of treatment, which suggest 430 431 that cerium is accumulated and transported by secondary roots from bottom towards the central part of radish roots, not only through the surface. Accumulation of the metals in the taproot of plants was 432 already observed after a LA-ICP-MS analysis of another root vegetables such as carrot (Yudasari et 433 al., 2018). In addition, two additional signals were observed for samples treated with  $CeO_2$  NPs for 434 435 two days, suggesting that after accumulation, nanoparticles can be translocated within analysed radish 436 tissues. Moreover, it was observed that the intensity of cerium signals increased with time for both 437 treatments, showing that a higher content of cerium can be accumulated in radish roots with a longer 438 exposure time; a relevant information in terms of food safety taking into account the persistence of 439 CeO<sub>2</sub> NPs in the environment. A good reproducibility of the cerium spatial distribution was achieved 440 after comparing results of the different slices of the same radish root or from different radish roots 441 with the same exposure concentration and time.



**443** Fig. 5. LA-ICP-MS analysis of radish roots treated with A) 5 mg  $L^{-1}$  CeO<sub>2</sub> NPs B) 50 mg  $L^{-1}$  CeO<sub>2</sub> 444 NPs. Light blue line: control plant; dark blue line: plants treated for 1 day; green line: plants treated 445 for 2 days

446

#### 447 **4.** Conclusions

The uptake, bioaccumulation, translocation, physico-chemical characterization and 448 localization of CeO<sub>2</sub> NPs in radish was investigated by using different mass spectrometry based 449 techniques. Our results showed that after cultivation, the majority of cerium remain in roots, with a 450 low transportation up to leaves and stems (TF 0.017) as it has been observed in other studies. In order 451 452 to get a correct particle size distribution, the behavior of CeO<sub>2</sub> NPs in plasma was investigated, showing that ICP-MS sensitivity depends on the physicochemical form of cerium and thus the 453 ionization efficiency must be taken into account. Although CeO<sub>2</sub> NPs underwent agglomeration in 454 hydroponic medium after 1 day, the SP-ICP-MS analysis indicated that radish only takes up small 455 nanoparticles, since the presence of bigger nanoparticles and/or aggregates was not observed. The 456 457 studies performed discarded interaction with plant matrix or effect of the enzymatic digestion procedure that affected the size of the nanoparticles. However, the analysis of leaves plus stems 458 459 showed the presence of nanoparticles with bigger sizes, which suggests that nanoparticles undergo agglomeration at the endpoint of their transportation. Finally, the analysis by LA-ICP-MS in radish 460 461 roots showed that the accumulation of  $CeO_2$  NPs occurs mainly below the skin surface but after accumulation they have the ability to enter and be translocated within the tissue. The results obtained 462 in this work can be consider meaningful from the point of view of food safety. 463

464

#### 465 Acknowledgements

466 This work was financially supported by the National Science Centre, Poland (grant n°
467 2015/18/M/ST4/00257) as well as by a STSM Mobility Grant (JW) COST Action TD1407 (grant n°
468 42007) and the project AQUITRACES Region Aquitaine 20131206001-13010973. Authors would

469 like to thank Wojciech Kurek (Department of Botany, Warsaw University of Life Sciences (SGGW))

470 for preparation of samples for LA-ICP-MS analysis.

471

#### 472 Appendix A. Supplementary data

- 473 Electronic supplementary information (ESI) available: procedure of extraction and fractionation of Ce
- 474 species from radish samples by SEC-ICP-MS; size distributions of CeO<sub>2</sub> NPs in grow medium at
- different incubation times; SEC-ICP-MS chromatogram of ammonium acetate extract of radish roots.
- 476

#### 477 **References**

- 478 Abad-Álvaro, I., Peña-Vázquez, E., Bolea, E., Bermejo-Barrera, P., Castillo, J.R., Laborda, F., 2016.
- Evaluation of number concentration quantification by single-particle inductively coupled plasma
  mass spectrometry: microsecond vs. millisecond dwell times. Anal. Bioanal. Chem. 408, 5089–
- 481 5097. https://doi.org/10.1007/s00216-016-9515-y
- Antisari, L.V., Carbone, S., Gatti, A., Vianello, G., Nannipieri, P., 2015. Uptake and translocation of
  metals and nutrients in tomato grown in soil polluted with metal oxide (CeO 2, Fe 3 O 4, SnO
  2, TiO 2) or metallic (Ag, Co, Ni) engineered nanoparticles. Env. Sci Pollut Res 1841–1853.
- 485 https://doi.org/10.1007/s11356-014-3509-0
- 486 Cassee, F.R., Balen, E.C. Van, Singh, C., Green, D., Muijser, H., 2011. Exposure , health and
- 487 ecological effects review of engineered nanoscale cerium and cerium oxide associated with its
  488 use as a fuel additive. Crit. Rev. Toxicol. 41, 213–229.
- 489 https://doi.org/10.3109/10408444.2010.529105
- Colvin, V.L., 2003. The potential environmental impact of engineered nanomaterials. Nat. Biotechnol.
  21, 1166–1170.
- 492 Dan, Y., Ma, X., Zhang, W., Liu, K., 2016. Single particle ICP-MS method development for the
  493 determination of plant uptake and accumulation of CeO 2 nanoparticles. Anal. Bioanal. Chem.

494 408, 5157–5167. https://doi.org/10.1007/s00216-016-9565-1

- Deng, Y., White, J.C., Xing, B., 2014. Interactions between engineered nanomaterials and agricultural
  crops : implications for food safety \*. J. Zhejiang Univ. Sci. A : Applied Phys. Eng. 15, 552–
- 497 572. https://doi.org/10.1631/jzus.A1400165
- 498 Gardea-torresdey, J.L., Rico, C.M., White, J.C., 2014. Trophic Transfer, Transformation, and Impact
- 499 of Engineered Nanomaterials in Terrestrial Environments. Environ. Sci. Technol. 48, 2529–
  500 2540. https://doi.org/10.1021/es4050665
- 501 Hernandez-viezcas, J.A., Castillo-michel, H., Andrews, J.C., Cotte, M., Rico, C., Peralta-videa, J.R.,

- 502 Ge, Y., Priester, J.H., Holden, A., Gardea-torresdey, J.L., 2013. In Situ Synchrotron X ray
- 503 Fluorescence Mapping and Speciation of CeO 2 and ZnO Nanoparticles in Soil Cultivated
- 504 Soybean (Glycine max). ACS Nano 7, 1415–1423. https://doi.org/10.1021/nn305196q
- 505 Hernandez-Viezcas, J.A., Castillo-michel, H., Peralta-Videa, J.R., Gardea-Torresdey, J.L., 2016.
- 506Interactions between CeO 2 Nanoparticles and the Desert Plant Mesquite: A Spectroscopy
- 507Approach. ACS Sustain. Chem. Eng. 4, 1187–1192.
- 508 https://doi.org/10.1021/acssuschemeng.5b01251
- Hu, S., Liu, R., Zhang, S., Huang, Z., Xing, Z., Zhang, X., 2009. A New Strategy for Highly Sensitive
  Immunoassay Based on Single-Particle Mode Detection by Inductively Coupled Plasma.
  https://doi.org/10.1016/j.jasms.2009.02.005
- 512 Jiménez-Lamana, J., Abad-Álvaro, I., Katarzyna, B., Laborda, F., Szpunar, J., Lobinski, R., 2018.
- 513 Detection and characterization of biogenic selenium nanoparticles in selenium-rich yeast by
- single particle ICPMS. J. Anal. At. Spectrom. 33, 452–460.
- 515 https://doi.org/10.1039/C7JA00378A
- 516 Jiménez-Lamana, J., Wojcieszek, J., Jakubiak, M., Asztemborska, M., Szpunar, J., 2016. Single
- 517 particle ICP-MS characterization of platinum nanoparticles uptake and bioaccumulation by
- 518 Lepidium sativum and Sinapis alba plants. J. Anal. At. Spectrom. 31, 2321–2329.
- 519 https://doi.org/10.1039/C6JA00201C
- Johnson, A.C., Park, B., 2012. Predicting contamination by the fuel additive cerium oxide engineered
   nanoparticles within the United Kingdom and the associated risks. Environ. Toxicol. Chem. 31,
   2582–2587. https://doi.org/10.1002/etc.1983
- Keller, A.A., Mcferran, S., Lazareva, A., Suh, S., 2013. Global life cycle releases of engineered
  nanomaterials. J Nanopart Res 15, 1692. https://doi.org/10.1007/s11051-013-1692-4
- Kińska, K., Jiménez-Lamana, J., Kowalska, J., Krasnodębska-Ostręga, B., Szpunar, J., 2018. Study of
   the uptake and bioaccumulation of palladium nanoparticles by Sinapis alba using single particle
- 527 ICP-MS. Sci. Total Environ. 615, 1078–1085. https://doi.org/10.1016/j.scitotenv.2017.09.203
- 528 Laborda, F., Jiménez-Lamana, J., Bolea, E., Castillo, J.R., 2011. Selective identification,
- 529 characterization and determination of dissolved silver(i) and silver nanoparticles based on single
- particle detection by inductively coupled plasma mass spectrometry. J. Anal. At. Spectrom. 26,
- 531 1362. https://doi.org/10.1039/c0ja00098a
- Li, Y., He, X., Yin, J., Ma, Y., Zhang, P., Li, J., Ding, Y., 2015. Acquired Superoxide-Scavenging
  Ability of Ceria Nanoparticles \*\*. Angew. Chemie Int. Ed. 54, 1832–1835.
- 534 https://doi.org/10.1002/anie.201410398
- 535 Lopez-Moreno, M.L., De la Rosa, G., Hernandez-Viezcas, J.A., Castillo-michel, H., Botez, C.E.,
- 536 Peralta-Videa, J.R., Gardea-Torresdey, J.L., 2010a. Evidence of the Differential
- 537 Biotransformation and Genotoxicity of ZnO and CeO 2 Nanoparticles on Soybean (Glycine
- 538 max ) Plants. Environ. Sci. Technol. 44, 7315–7320.

- 539 Lopez-Moreno, M.L., De la Rosa, G., Hernandez-Viezcas, J.A., Peralta-Videa, J.R., Gardea-
- 540 Torresday, J.L., 2010b. X-ray Absorption Spectroscopy (XAS) Corroboration of the Uptake
  541 and Storage of CeO 2 Nanoparticles and Assessment of Their Differential Toxicity in Four
- Edible Plant Species. J. Agric. Food Chem. 58, 3689–3693. https://doi.org/10.1021/jf904472e
- 543 Ma, X., Geiser-lee, J., Deng, Y., Kolmakov, A., 2010. Interactions between engineered nanoparticles (
- 544 ENPs ) and plants : Phytotoxicity , uptake and accumulation. Sci. Total Environ. 408, 3053–
- 545 3061. https://doi.org/10.1016/j.scitotenv.2010.03.031
- Ma, X., Wang, Q., Rossi, L., Ebbs, S.D., White, J.C., 2016. Multigenerational exposure to cerium
  oxide nanoparticles : Physiological and biochemical analysis reveals transmissible changes in
  rapid cycling Brassica rapa. NanoImpact 1, 46–54. https://doi.org/10.1016/j.impact.2016.04.001
- 549 Ma, Y., Zhang, P., Zhang, Z., He, X., Zhang, Junzhe, Ding, Y., Zhang, Jing, Zheng, L., Guo, Z.,
- 550 Zhang, L., Chai, Z., Zhao, Y., 2015. Where Does the Transformation of Precipitated Ceria
- 551 Nanoparticles in Hydroponic Plants Take Place? Environ. Sci. Technol. 9, 10667–10674.
- 552 https://doi.org/10.1021/acs.est.5b02761
- Majumdar, S., Peralta-videa, J.R., Bandyopadhyay, S., 2014. Exposure of cerium oxide nanoparticles
  to kidney bean shows disturbance in the plant defense mechanisms. J. Hazard. Mater. 278, 279–
  287. https://doi.org/10.1016/j.jhazmat.2014.06.009
- 556 Majumdar, S., Peralta-videa, J.R., Trujillo-reyes, J., Sun, Y., Barrios, A.C., Niu, G., Margez, J.P.F.-,
- 557 Gardea-torresdey, J.L., 2016. Science of the Total Environment Soil organic matter in fl uences
- 558 cerium translocation and physiological processes in kidney bean plants exposed to cerium oxide
- nanoparticles. Sci. Total Environ. 569–570, 201–211.
- 560 https://doi.org/10.1016/j.scitotenv.2016.06.087
- 561 Pace, H.E., Rogers, N.J., Jarolimek, C., Coleman, V.A., Higgins, C.P., Ranville, J.F., 2011.
- 562 Determining Transport Efficiency for the Purpose of Counting and Sizing Nanoparticles via
- 563 Single Particle Inductively Coupled Plasma Mass Spectrometry. Anal. Chem. 83, 9361–9369.
  564 https://doi.org/10.1021/ac201952t
- Piccinno, F., Gottschalk, F., Seeger, S., Nowack, B., 2012. Industrial production quantities and uses of
  ten engineered nanomaterials in Europe and the world. J Nanopart Res 14, 1109.
- 567 https://doi.org/10.1007/s11051-012-1109-9
- Rossi, L., Zhang, W., Lombardini, L., Ma, X., 2016. The impact of cerium oxide nanoparticles on the
  salt stress responses of Brassica napus L . \*. Environ. Pollut. 219, 28–36.
- 570 https://doi.org/10.1016/j.envpol.2016.09.060
- 571 Sánchez-garcía, L., Bolea, E., Laborda, F., Cubel, C., Ferrer, P., Gianolio, D., Silva, I., Castillo, J.R.,
- 572 2016. Size determination and quantification of engineered cerium oxide nanoparticles by flow
- 573 field-flow fractionation coupled to inductively coupled plasma mass spectrometry. J.
- 574 Chromatogr. A 1438, 205–215. https://doi.org/10.1016/j.chroma.2016.02.036
- 575 Schwabe, F., Schulin, R., Limbach, L.K., Stark, W., Bürge, D., Nowack, B., 2013. Influence of two

- 576 types of organic matter on interaction of CeO 2 nanoparticles with plants in hydroponic culture.
- 577 Chemosphere 91, 512–520. https://doi.org/10.1016/j.chemosphere.2012.12.025
- Schwabe, F., Schulin, R., Rupper, P., Rotzetter, A., Stark, W., Nowack, B., 2014. Dissolution and
  transformation of cerium oxide nanoparticles in plant growth media. J. Nanoparticle Res. 16,
  2668. https://doi.org/10.1007/s11051-014-2668-8
- 581 Schwabe, F., Tanner, S., Schulin, R., Rotzetter, A., Stark, W., Quadt, A. Von, Nowack, B., 2015.
- 582 Dissolved cerium contributes to uptake of Ce in the three crop plants †. Metallomics 7, 466–477.
  583 https://doi.org/10.1039/C4MT00343H
- Stanek, C.R., Tan, A.H.H., Owens, S.L., Grimes, Æ.R.W., 2008. Atomistic simulation of CeO 2
  surface hydroxylation : implications for glass polishing. J Mater Sci 43, 4157–4162.
  https://doi.org/10.1007/s10853-008-2605-2
- Stark, W.J., Stoessel, P.R., Wohlleben, W., Hafner, A., 2015. Industrial applications of nanoparticles.
  Chem Soc Rev 44, 5793–5805. https://doi.org/10.1039/c4cs00362d
- 589 Vance, M.E., Kuiken, T., Vejerano, E.P., McGinnis, S.P., Hochella, M.F., Hull, D.R., 2015.
- 590 Nanotechnology in the real world: Redeveloping the nanomaterial consumer products inventory.
  591 Beilstein J. Nanotechnol. 6, 1769–1780. https://doi.org/10.3762/bjnano.6.181
- 592 Walser, T., Limbach, L.K., Brogioli, R., Erismann, E., Flamigni, L., Hattendorf, B., Juchli, M.,
- 593 Krumeich, F., Ludwig, C., Prikopsky, K., Rossier, M., Saner, D., Sigg, A., Hellweg, S., Gu, D.,
- 594 Stark, W.J., 2012. Persistence of engineered nanoparticles in a municipal solid-waste
- incineration plant. Nat. Nanotechnol. 7, 6–11. https://doi.org/10.1038/NNANO.2012.64
- Wang, Q., Ma, X., Zhang, W., Peia, H., Chen, Y., 2012. The impact of cerium oxide nanoparticles on
  tomato (Solanum lycopersicum L .) and its implications for food safety. Metallomics 4, 1105–
  1112. https://doi.org/10.1039/c2mt20149f
- Wojcieszek, J., Szpunar, J., Lobinski, R., 2018. Speciation of technologically critical elements in the
  environment using chromatography with element and molecule specific detection. Trends Anal.
  Chem. 104, 42–53. https://doi.org/10.1016/j.trac.2017.09.018
- Yang, X., Pan, H., Wang, P., Zhao, F., 2017. Particle-specific toxicity and bioavailability of cerium
  oxide (CeO 2) nanoparticles to Arabidopsis thaliana. J. Hazard. Mater. 322, 292–300.
  https://doi.org/10.1016/j.jhazmat.2016.03.054
- 605 Yokel, R.A., Florence, R.L., Unrine, J.M., Tseng, M.T., Graham, U.M., Wu, P., Grulke, E.A.,
- 606 Sultana, R., Hardas, S.S., Butterfield, D.A., 2009. Biodistribution and oxidative stress effects of
- a systemically-introduced commercial ceria engineered nanomaterial. Nanotechnology 3, 234–
  249. https://doi.org/10.1080/17435390902974496
- 609 Yudasari, Y., Prasetyo, S., Suliyanti, M.M., 2018. The 1064 nm laser-induced breakdown
- spectroscopy (LIBS) inspection to detect the nutrient elements in freshly cut carrot samples. J.
  Phys. Conf. Ser 985, 1–6.
- 612 Zhang, P., Ma, Y., Liu, S., Wang, G., Zhang, Junzhe, He, X., Zhang, Jing, Rui, Y., Zhang, Z., 2017a.

- 613 Phytotoxicity, uptake and transformation of nano-CeO 2 in sand cultured romaine lettuce \*.
- 614 Environ. Pollut. 220, 1400–1408. https://doi.org/10.1016/j.envpol.2016.10.094
- 615 Zhang, P., Ma, Y., Zhang, Z., He, X., Zhang, J., Guo, Z., Tai, R., Zhao, Y., Chai, Z., 2012.
- Biotransformation of Ceria Nanoparticles in Cucumber Plants. ACS Nano 9943–9950.
  https://doi.org/10.1021/nn303543n
- 618 Zhang, P., Xie, C., He, X., Zhang, Z., Ding, Y., Zheng, L., Zhang, J., 2017b. Shape-Dependent
- Transformation and Translocation of Ceria Nanoparticles in Cucumber Plants. Environ. Sci.
  Technol. Lett. 4, 380–385. https://doi.org/10.1021/acs.estlett.7b00359
- Zhang, W., Dan, Y., Shi, H., Ma, X., 2017a. Elucidating the mechanisms for plant uptake and inplanta speciation of cerium in radish (Raphanus sativus L.) treated with cerium oxide
- 623 nanoparticles. J. Environ. Chem. Eng. 5, 572–577. https://doi.org/10.1016/j.jece.2016.12.036
- 624 Zhang, W., Ebbs, S.D., Musante, C., White, J.C., Gao, C., Ma, X., 2015. Uptake and Accumulation of
- Bulk and Nanosized Cerium Oxide Particles and Ionic Cerium by Radish (Raphanus sativus L.).
- 626 J. Agric. Food Chem. 63, 382–390. https://doi.org/10.1021/jf5052442
- 627 Zhang, W., Musante, C., White, J.C., Schwab, P., Wang, Q., Ebbs, S.D., Ma, X., 2017b.
- Bioavailability of cerium oxide nanoparticles to Raphanus sativus L. in two soils. Plant Physiol.
  Biochem. 110, 185–193. https://doi.org/10.1016/j.plaphy.2015.12.013
- Zhang, Z., He, X., Zhang, H., Ma, Y., Zhang, P., Ding, Y., Zhao, Y., 2011. Uptake and distribution of
  ceria nanoparticles in cucumber plants. Metallomics 3, 816–822.
- 632 https://doi.org/10.1039/c1mt00049g
- 633 Zhao, L., Peralta-videa, J.R., Rico, C.M., Hernandez-viezcas, J.A., Sun, Y., Niu, G., Servin, A.,
- 634 Nunez, J.E., Duarte-gardea, M., Gardea-torresdey, J.L., 2014. CeO 2 and ZnO Nanoparticles
- 635 Change the Nutritional Qualities of Cucumber ( Cucumis sativus ). J. Agric. Food Chem. 62,
- 636 2752–2759. https://doi.org/10.1021/jf405476u
- 637 Zhao, L., Sun, Y., Hernandez-Viezcas, J.A., Servin, A.D., Hong, J., Niu, G., Peralta-Videa, J.R.,
- 638 Duarte-Gardea, M., Gardea-Torresdey, J.L., 2013. Influence of CeO2 and ZnO nanoparticles on
- 639 cucumber physiological markers and bioaccumulation of Ce and Zn: A Life Cycle Study. J.
- 640 Agric. Food Chem. 61, 11945–11951. https://doi.org/10.1021/jf404328e
- 641

